# Toxics Chemicals in NYS Tributaries to Lake Ontario: A Report on Sampling Undertaken in 2007 and 2008 with Special Emphasis on the Polychlorinated Dibenzodioxins and Furans 



Report to the U.S. Environmental Protection Agency, Fred Luckey, Project Manager

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## Summary

This project was undertaken to provide estimates of loading of synthetic chemicals into Lake Ontario from several New York tributaries.

We estimate that 18-Mile Creek, Genesee River, Oswego River, Salmon River, and Black River together contribute $97 \mathrm{~g} /$ day total mercury, $38 \mathrm{~g} /$ day dissolved mercury, $66 \mathrm{~g} /$ day PCB, and $2.5 \mathrm{mg} /$ day $2,3,7,8-\mathrm{TCDD}$ equivalents to Lake Ontario.

The highest mercury loading rates ( $\mathrm{g} /$ day, $\mathrm{g} /$ capita, and $\mathrm{g} / \mathrm{sq} \mathrm{km}$ ) were seen in the Black River. PCB loading rates were particularly high on 18-Mile Creek. 18-Mile Creek also showed high concentrations of pesticides, particularly DDT and its metabolites. 18-Mile Creek showed very high dioxin dioxin loads on an area basis but less so on a per capita basis. Population density in a drainage basin may turn out to be an important predictor of dioxin concentrations.

The PCDD/F congener patterns in Lake Ontario sediments appear to be more like those of the Niagara River sites at Cayuga Island and Love Canal than of the principal tributaries.

The Erie Canal may move PCDD/Fs from the Tonawanda/Lockport area to the Genesee River. Overall, congeners 2,3,4,7,8-PeCDF (congener 10), 1,2,3,4,6,7,8-HpCDD (congener 6), and $1,2,3,7,8-\mathrm{PeCDD}$ (congener 2) are most often the greatest contributors to TEQ in the water samples. Congeners 10 and 2 are often associated with combustion. We have not been able to identify industrial processes that generate congener 6.

Raw concentrations of congener 6 are in most samples about a tenth that of octachlorodioxin (congener 7). Mono-dechlorination of congener 7 can produce only $1,2,3,4,6,7,9-\mathrm{HpCDD}$ (which is not reported in regulatory work) or congener 6. Congener 6 has a TEF 100 times that of congener 7 and its bioaccumulation factor is 5 times greater than that of congener 7. Microbial dechlorination of congener 7 does occur in laboratory experiments. [1] Dechlorination of the very large reservoir of octachlorodioxin could be a long term source of dioxin toxicity in NYS sediments.

## Introduction

The New York State Department of Health recommends that women of childbearing age and children under the age of 15 eat no freshwater fish. Levels of PCBs, Mirex, and dioxins are problematic in many fish species (causing consumption recommendations) throughout Lake Ontario, from the Niagara River, and from 18- Mile Creek in Niagara County.[2] The sources of Lake Ontario contaminants are from historic sediment deposition, mass transfer from upstream water (assessed by the Upstream/Downstream Niagara River Monitoring Program), the Lake Ontario Air Deposition Study, point sources (assessed in NYS through the State Pollution Discharge Elimination System SPDES), and tributary inputs. Tributary inputs have been assessed in the past (from 2002-2008) in the NYS tributaries by Richard Coleates of USEPA Region 2 and, for the Black River, by Richard and Eckhardt.[3].

Table 1. Tributary data from USEPA Region 2.

| site | date | $\begin{gathered} \text { Total DDT } \\ \mathrm{ng} / \mathrm{L} \\ \hline \end{gathered}$ | Dieldrin ng/L | Total Mercury ng/L | $\begin{gathered} \text { PCBs } \\ \mathrm{ng} / \mathrm{L} \end{gathered}$ | $\begin{gathered} \text { Dioxins } \\ \text { TEQ pg/L } \\ \hline \end{gathered}$ |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 18-Mile Creek | 4/16/02 | U | U | 12.4 | 35.7 | U |
| 18-Mile Creek | 9/17/02 | U | U | 0.863 | 32.5 | 13.9 |
| 18-Mile Creek | 5/6/03 | U | U | 4.53 | 29.6 | 0.016 |
| 18-Mile Creek | 7/9/03 | U | U | 1.43 | 38.7 | U |
| 18-Mile Creek | 10/7/03 | U | U | 1.3 | 21.5 | U |
| 18-Mile Creek | 5/11/04 | U | U | 4.6 | 51.3 | NA |
| 18-Mile Creek | 9/28/04 | U | U | 1.35 | 39.5 | NA |
| 18-Mile Creek | 5/3/05 | 0.943 | 0.276 | 3.28 | 35.5 | NA |
| 18-Mile Creek | 8/30/05 | 0.807 | 0.375 | 2.07 | 47.3 | NA |
| 18-Mile Creek | 7/26/06 | NA | NA | 1.42 | 50.4 | NA |
| 18-Mile Creek | 9/19/06 | NA | NA | 5.73 | 52.2 | NA |
| 18-Mile Creek | 6/26/07 | NA | NA | 1.03 |  | NA |
| 18-Mile Creek | 10/16/08 | NA | NA | 1.22 |  | NA |
|  | median | 0.875 | 0.3255 | 1.43 | 38.7 | 6.958 |
| Black River | 4/18/02 | U | U | 4.99 | 1.85 | U |
| Black River | 9/18/02 | U | U | 1.67 | 0.76 | U |
| Black River | 5/7/03 | U | U | 3.55 | 0.425 | NA |
| Black River | 7/10/03 | U | U | 2.5 | 1.17 | NA |
| Black River | 10/8/03 | U | U | 4.65 | 0.417 | NA |
| Black River | 5/12/04 | U | U | 2.74 | 1.31 | NA |
| Black River | 9/29/04 | U | U | 2.46 | 19.5 | NA |
| Black River | 5/5/05 | QB | U | 2.82 | 12.2 | NA |
| Black River | 7/25/05 | NA | NA | 2.84 | 1.52 | NA |
| Black River | 9/1/05 | U | U | 5.55 | 10.3 | NA |
| Black River | 9/20/06 | NA | NA | 2.19 | 0.385 | NA |
| Black River | 6/27/07 | NA | NA | 2.15 |  | NA |
| Black River | 10/17/08 | NA | NA | 2.57 |  | NA |
|  | median |  |  | 2.74 | 1.31 |  |
| Genesee R. | 4/16/02 | U | U | 10.9 | 0.157 | 0.041 |
| Genesee R. | 9/17/02 | U | U | 1.13 | 0.414 | U |
| Genesee R. | 5/6/03 | U | U | 2.26 | U | U |
| Genesee R. | 7/9/03 | U | U | 1.83 | 0.015 | U |
| Genesee R. | 10/7/03 | U | U | 1.97 | 0.256 | U |
| Genesee R. | 5/11/04 | U | U | 2.53 | 0.022 | NA |
| Genesee R. | 9/28/04 | U | U | 4.23 | 0.149 | NA |
| Genesee R. | 5/4/05 | QB | U | 2.63 | 0.313 | NA |

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Table 1. Continued.

| site | date | Total DDT ng/L | Dieldrin ng/L | Total Mercury ng/L | PCBs <br> ng/L | Dioxins TEQ pg/L |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Genesee R. | 8/30/05 | U | U | 1.14 | 0.338 | NA |
| Genesee R. | 7/26/06 | NA | NA | 3.16 | 0.358 | NA |
| Genesee R. | 9/20/06 | NA | NA | 2.81 | 0.596 | NA |
| Genesee R. | 6/26/07 | NA | NA | 1.18 |  | NA |
| Genesee R. | 10/16/08 | NA | NA | 2.15 |  | NA |
|  | median |  |  | 2.26 | 0.2845 |  |
| Oswego R. | 4/17/02 | U | U | 3.31 | 0.166 | U |
| Oswego R. | 9/18/02 | U | U | 1.24 | 0.366 | U |
| Oswego R. | 5/7/03 | U | U | 1.59 | U | NA |
| Oswego R. | 7/10/03 | U | U | 1.25 | 0.017 | NA |
| Oswego R. | 10/8/03 | U | U | U | 0.203 | NA |
| Oswego R. | 5/12/04 | U | U | 2.2 | 0.193 | NA |
| Oswego R. | 9/29/04 | U | U | 1.3 | 0.54 | NA |
| Oswego R. | 5/4/05 | QB | U | 1.71 | 4.76 | NA |
| Oswego R. | 7/25/05 | NA | NA | 1.69 | 0.335 | NA |
| Oswego R. | 8/30/05 | U | U | U | 0.107 | NA |
| Oswego R. | 9/20/06 | NA | NA | 1.22 | 0.026 | NA |
| Oswego R. | 6/27/07 | NA | NA | 0.869 |  | NA |
| Oswego R. | 10/17/08 | NA | NA | 0.78 |  | NA |
|  | median |  |  | 1.3 | 0.198 |  |
| Salmon R. | 4/17/02 | U | U | 2.85 | 0.3 | U |
| Salmon R. | 9/18/02 | U | U | 0.915 | 0.257 | U |
| Salmon R. | 5/7/03 | U | U | 2.18 | U | NA |
| Salmon R. | 7/10/03 | U | U | 1.68 | 0.013 | NA |
| Salmon R. | 10/8/03 | U | U | 1.92 | 0.149 | NA |
| Salmon R. | 5/12/04 | U | U | 2.22 | U | NA |
| Salmon R. | 9/29/04 | U | U | 1.74 | 0.473 | NA |
| Salmon R. | 5/5/05 | QB | U | 1.68 | 7.4 | NA |
| Salmon R. | 7/25/05 | NA | NA | 1.95 | U | NA |
| Salmon R. | 9/1/05 | U | U | 1.178 | 0.848 | NA |
| Salmon R. | 9/20/06 | NA | NA | 1.83 | 0.39 | NA |
| Salmon R. | 6/27/07 | NA | NA | 1.29 |  | NA |
| Salmon R. | 10/17/08 | NA | NA | 0.868 |  | NA |
|  | median |  |  | 1.74 | 0.345 |  |

Each project has its own methods, sampling sites, and approaches to data handling.
Here we will look at some recent data on mercury (total and dissolved), polychlorinated biphenyls (PCBs), chlorinated pesticides, and, most intensely, polychlorinated dibenzo-pdioxins and polychlorinated dibenzofurams (PCDD/Fs) from five New York State Tributaries - 18- Mile Creek, Genesee River, Oswego River, Salmon River, and the Black River.

## The Watersheds



Figure 1. Sampling sites visited and watersheds.
The sampling locations were selected to be as near Lake Ontario as possible while being upstream of lake influence, safe, and practical. With the exception of Lock 8 at Oswego, all the samples were taken from bridges. We were unable to perform equal area or equal discharge-type sampling due to the requirement of obtaining very large volume samples for PCDD/F analysis.

TABLE 2. Sampling Sites

| site name |  |  | sq km <br> area |  |  | 2000 <br> pop |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: |
| sampled |  |  |  |  |  |  |$|$

*Population is only from the NY portion of the watershed.
Discharges from the Genesee, Oswego, Black Rivers were taken from U.S. Geological Survey (USGS) gages that were either at the sampling location or very near. The Salmon River is gauged at Pineville and 18-Mile Creek is not gauged. To obtain discharges for Salmon River at Pulaski the USGS discharge data from Pineville were multiplied by the ratio:
[watershed area above Pulaski]/[watershed area above Pineville] = 1.157

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18- Mile Creek, as noted, is not gauged. Discharge from near-by Tonawanda Creek at Rapids, NY was multiplied by the ratio:
[watershed area of 18-Mile Cr. above Jacques Rd.]/[watershed area of Tonawanda Cr. above Rapids, NY] $=0.205$.

18-Mile Creek receives 50 cfs in overflow from the Erie Canal at Lockport and 15 cfs from overflow of the Erie Canal at Gasport during canal season (May 1 to November 15). It also receives discharge from the Lockport POTW. Daily average POTW discharges were added.

The study period was $5 / 1 / 2007$ to $10 / 1 / 2008$. Table 3 shows discharge statistics from the sampled streams during the study period and statistics of discharges during sampling events.

TABLE 3. Discharge (cubic feet per second) statistics for period of record and for sampling events.

| $5 / 1 / 2007-10 / 1 / 2008$ | Black | Oswego | Genesee | Salmon | 18 Mile Cr. |
| :--- | :---: | :---: | :---: | :---: | :---: |
| mean | 4,578 | 6,497 | 2,434 | 836 | 132 |
| median | 3,240 | 3,580 | 1,520 | 577 | 97 |
| $50 \%$ disch. $^{1}$ | $20.60 \%$ | $20.00 \%$ | $18.30 \%$ | $22.10 \%$ | $26.92 \%$ |
| Sampling events $^{\text {mean }}$ | 6,188 | 12,276 | 3,214 | 2,363 | 176 |
| percentile $^{2}$ | 77 | 80 | 78 | 95 | 85 |
| median $^{5,065}$ | 11,900 | 1,691 | 2,400 | 158 |  |

1. Half of the total flow in the Black R. occurred over $20.6 \%$ of the time
2. The mean discharge in the Black $R$. during sampling events was at the $77^{\text {th }}$ percentile of daily discharges in the Black over the study period.

## Analysis

Table 4 summarizes the projects parameters, procedures, preservatives, holding times, detection limits, precision, accuracy, and methods for organic analytes.

TABLE 4. Analytical methods.

|  | PCDD/F | PCB | Pesticides | POC | Hg |
| :---: | :---: | :---: | :---: | :---: | :---: |
| Procedure | filter | grab | Grab | filtration | grab/filtration |
| Preservative | $<=4^{\circ} \mathrm{C}$ | <= $4^{\circ} \mathrm{C}$ | < $=4^{\circ} \mathrm{C}$ | freezing | $<=4^{\circ} \mathrm{C}$ |
| Holding time | 1 year | 7 days | 7 days | 6 months | 7 days |
| Detection Limit | $0.091 \mathrm{pg} / \mathrm{L}$ | 0.098 pg/L | $42 \mathrm{pg} / \mathrm{L}$ (Total DDT) | $0.01 \mathrm{mg} / \mathrm{L}$ | $0.2 \mathrm{ng} / \mathrm{L}$ |
| Accuracy (RSD) | 25-164\% | 50-150\% | 76-116\% |  | 77-123\% |
| Precision | 27\% | 40\% | 21\% | 5\% | 8.30\% |
| Field/Lab Method | 1613B | 1668A | NYSDECHRMS2 | wet oxidation | 1631 |

## Sampling

Sampling was conducted using a modification of the Trace Organics Platform Sampler (TOPS) procedure. TOPS is specifically designed to field-concentrate highly dilute hydrophobic chemicals but the procedure facilitates other kinds of sampling. Pressurized
water can be pushed through a filter to obtain dissolved mercury samples, water is readily available for filtering particulate carbon, and water is available for collecting whole-water PCB/pesticide samples. Figures 2 and 3 schematically illustrate the TOPS set-up.

## Sampling for PCDD/Fs (Dioxins)

Dioxins were sampled by pumping water through pre-cleaned glass fiber cartridge filters having a nominal porosity of $1 \mu \mathrm{~m}$. The system is run for a minimum of 10 minutes (timed with a stop watch) before the cartridge filter is mounted in the stainless steel housing. Filtered water is wasted. The filters were previously ashed 4 hours at $450^{\circ} \mathrm{C}$, wrapped in aluminum foil, double bagged, and stored prior to use in a laboratory freezer. One (or two, depending on the distance from the water to the bridge deck) 5C-MD March magnetic impeller pump brings water from a stainless steel intake mounted on an epoxy painted sampling fish suspended 2 feet below the water's surface. The shaped "fish" orients into the current. Water comes to the deck of the bridge where it enters a tee. Some of the water is wasted (used for other aspects of sampling) and some is drawn by a peristaltic pump into a TOPS. Inside the TOPS, the water passes a pressure sensor and then goes through a cartridge filter held in a stainless steel housing. Sampled water is not exposed to the air and is completely self-contained.

The pump rate is usually held constant and is measured by noting the time required to fill a 20 L plastic carboy. The carboy is weighed on a certified scale full and empty. The flow rate is calculated. During each deployment carboys are filled and weighed at least three times. If field conditions are unfavorable, the carboys can be capped and brought back to the lab for weighing. If the flow rate is changed, the flows are re-calibrated. Careful note is taken of start and stop times for the pump. To obtain the volume of water filtered the average pump rate ( $\mathrm{L} / \mathrm{min}$ ) is multiplied by the minutes of pumping.

When the pressure sensor detects a back-pressure of 15 psi the TOPS shuts off. Ideally, TOPS is run to automatic shut-off but in practice, this is not always achievable. The stainless steel housing is opened and the loaded glass fiber cartridge (and any residual water) is quickly dumped into a certified wide-mouth bottle which had been previously rinsed 3X with site water. The bottle is capped, labeled, and brought back to a refrigerator at the NYSDEC laboratory in Albany prior to being shipped out for analysis at a contract analytical lab. Between field deployments the TOPS is cleaned by recirculating hot soapy water for 10 minutes. As much soapy water as possible is drained before rinsing the unit with fresh hot water for 10 minutes. Two blanks were created by loading clean glass fiber cartridges into the TOPS after cleaning and pumping through 4 L of nanograde laboratory water. A field duplicate was created by running two TOPS units simultaneously off a single intake.


Figure 2. Set-up for bringing water up from a stream channel, collecting samples, and filtration.

Items within the box are called TOPS.

Figure 3. Set up for collecting filtered water for dissolved mercury.

## Sampling for Mercury.

Water coming up from the tow fish passes through a tee (part D). Tubing to the right of part A is polyethylene. Some water goes to the right and some is wasted (going to the left in the illustration). Mercury sampling was done with a modified clean hands procedure after the system has been flushed with site water for about one hour. Filters and bottles are supplied by the lab double bagged in Ziplocs. The person playing the role of "Dirty Hands", while wearing Class 100 gloves, assists Clean Hands in putting on his Class 100 gloves. During the entire process Clean Hands touches nothing other than the inner surfaces of the outer Ziploc, the inner Ziploc, and the bottles and filters. Dirty Hands opens the outer Ziploc bag holding part A -a Gelman Sciences $0.45 \mu \mathrm{~m}$ high capacity groundwater sampling capsule. Clean Hands opens the inner Ziploc and extracts the capsule. Dirty Hands steadies the $1 / 2$ inch ID polyethylene tube (to the left of the tee)
while Clean Hands inserts the intake end of the sampling capsule into the polyethylene. Filtered water now exits through tube B which has been pre-cleaned and was attached to the sampling capsule by the mercury analytical lab prior to double bagging. Clean Hands makes sure to not let the tube B contact foreign surfaces. After a few minutes of flushing, Dirty and Clean Hands have un-bagged a sample container (part C). Clean Hands shields the opening of the sample bottle while it is being filled to prevent dust from entering. After filling the sample bottle is quickly capped and re-inserted into the double bags. The process is repeated without the filter for the whole water sample. Clean Hands takes care not to contact the sample bottle with the polyethylene tubing. Sample labels are attached to the inner surface of the outer Ziploc.

## Sampling for Particulate Organic Carbon

Particulate organic carbon (POC) was measured in raw and filtered (post TOPS) water. The glass fiber cartridge filters used in TOPS have a high capacity for suspended particles but they are fairly coarse (nominal porosity of $1 \mu \mathrm{~m}$ ) and are thus inefficient. Trapping efficiency is a function of particle size and particle loading. As filters load, porosity decreases making the filter more efficient. We measure TOPS trapping efficiency by using $0.7 \mu \mathrm{~m}$ flat glass fiber GFF filters. The GFF filters were partially wrapped in aluminum foil and ashed at $450^{\circ} \mathrm{C}$ for 4 hours; resealed while warm; and stored doubled bagged in Ziplocks in a laboratory freezer prior to deployment.

Broad-headed stainless steel forceps are used to place the filters into a magnetic filter holder set into a 2 L side arm Erlenmeyer flask. Vacuum is supplied by an electric bench-top pump. The magnetic filter holder has a 200 mL reservoir. Prior to loading the filter, the assembly is rinsed with site water. Water that had passed through the TOPS glass fiber cartridge filter is processed first. Water is filtered until plugging reduces flow to a drip. The reservoir is fully filtered and the vacuum is broken. The broad-head forceps are used to transfer the filter into a clean close fitting plastic Petri dish. The resealed Petri is labeled and kept cold in the field and frozen in the lab prior to analysis. The filtering process is repeated on for raw water.

## Sampling for PCBs and Pesticides

Whole water samples for PCBs and pesticides were obtained by filling 1 L pre-clean certified amber bottles (Boston Rounds) with water from the same $1 / 2$ inch polyethylene line as was used for the mercury samples. Three bottles were filled for each sample. The bottles were rinsed 3 x with site water prior to final filling.

## Mercury Results

Table 5 gives mercury concentrations and loads. Highlighted values were less than 5 x the maximum (of 3) field blanks. There were no field blanks for filtered water.

Mercury loads were estimated for the entire period of record by calculating a log concentration - log discharge linear relationship. The slope and intercept coefficients were applied to the known or (in the case of Salmon and 18-Mile Creek) calculated discharges so that a mercury concentration was estimated for each day. In cases where there were missing records, the mean between the days preceding and following the censored observations was applied.

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TABLE 5. Mercury samples, results, and instantaneous loads.

| site | date | discharge cf/sec | g/hour filtered | g/hour whole | ng/L <br> filtered | ng/L <br> whole |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 18-Mile Cr. | 5/1/2007 | 210 | 0.02 | 0.06 | 1.11 | 3.35 |
|  | 3/3/2008 | 101 | 0.01 | 0.14 | 2.49 | 29.8 ${ }^{\text {\# }}$ |
|  | 3/24/2008 | 273 | 0.07 | 0.19 | 2.86 | 7.42 |
|  | 4/7/2008 | 301 | 0.12 | 0.15 | 4.28 | 5.16 |
|  | 4/29/2008 | 70 | 0.02 | 0.03 | 1.39 | 2.33 |
|  | 5/27/2008 | 107 | 0.01 | 0.05 | 0.56 | 5.41 |
|  | 6/9/2008 | 96 | 0.01 | 0.07 | 1.73 | 8.42 |
| 18-Mile Cr. average |  | 165 | 0.037 | 0.099 | 2.06 | 5.35 |
| Black R. | 5/3/2007 | 8730 | 1.59 | 2.5 | 1.79 | 2.81 |
|  | 1/9/2008 | 14000 | 3.64 | 14.56 | 2.55 | 10.2* |
|  | 3/13/2008 | 12200 | 2.13 | 3.62 | 1.71 | 2.91 |
|  | 5/28/2008 | 2660 | 0.28 | 0.48 | 1.04 | 1.78 |
|  | 6/10/2008 | 3580 | 0.53 | 1.42 | 1.45 | 3.88 |
| Black R. average |  | 8234 | 1.63 | 4.52 | 1.71 | 2.85 |
| Genesee R. | 5/2/2007 | 4250 | 0.29 | 1.15 | 0.66 | 2.65 |
|  | 3/4/2008 | 4680 | 0.52 | 1.97 | 1.09 | 4.13 |
|  | 4/8/2008 | 8030 | 0.76 | 3.5 | 0.93 | 4.27 |
|  | 5/28/2008 | 1680 | 0.1 | 0.31 | 0.56 | 1.79 |
|  | 6/10/2008 | 1417 | 0.04 | 0.18 | 0.28 | 1.24 |
|  | 6/19/2008 | 827 | 0.05 | 0.16 | 0.65 | 1.9 |
|  | 9/30/2008 | 1451 |  | 0.15 | <0.15 | 1.01 |
| Genesee R. average |  | 3191 | 0.293 | 1.06 | 0.695 | 2.43 |
| Oswego R. | 5/2/2007 | 19700 | 1.53 | 4.9 | 0.76 | 2.44 |
|  | 10/25/2007 | 9570 | 0.33 | 2.38 | 0.34 | 2.44 |
|  | 1/10/2008 | 15200 | 1.32 | 2.96 | 0.85 | 1.91 |
|  | 3/4/2008 | 11900 | 0.76 | 1.33 | 0.63 | 1.1 |
|  | 3/25/2008 | 16600 | 2.47 | 4.48 | 1.46 | 2.65 |
|  | 4/30/2008 | 11800 | 1.72 | 3.06 | 1.43 | 2.54 |
|  | 9/30/2008 | 1160 | 0.03 | 0.08 | 0.28 | 0.64 |
| Oswego R. average |  | 12276 | 1.17 | 2.74 | 0.821 | 1.96 |
| Salmon R. | 5/3/2007 | 1533 | 0.23 | 0.34 | 1.48 | 2.19 |
|  | 1/10/2008 | 2581 | 0.47 | 0.69 | 1.8 | 2.63 |
|  | 4/8/2008 | 2998 | 0.45 | 0.88 | 1.47 | 2.89 |
|  | 5/1/2008 | 2096 | 0.4 | 0.52 | 1.85 | 2.43 |
| Salmon R. average |  | 2302 | 0.390 | 0.610 | 1.65 | 2.54 |

\# value not used in load calculations.
1/9/2008 was the occasion of a significant weather anomaly. Temperatures in much of the eastern US were $30^{\circ} \mathrm{F}$ above normal. Wind gusts at Watertown reaching 60 mph on the day of sampling made operations extremely difficult. The high mercury concentration may have been the result of wind blown dust.

Correlations between concentration and discharge were poor for total and dissolved mercury in 18-Mile Creek and poor for dissolved mercury at Salmon River. These data suggest that the Black River is an important mercury source to Lake Ontario. Salmon R. shows a higher mercury loading rate per sq km and per capita than do the Genesee or the Oswego Rivers (Table 6).

TABLE 6. Mercury loads derived from log/log regressions.

|  |  |  |  |  |  |  |  |
| :--- | ---: | ---: | ---: | :---: | ---: | ---: | ---: |
| total mercury |  | slope | intercept | r2 | total load (g) | g/sq km | g/capita |
| g/day |  |  |  |  |  |  |  |
| 18-Mile Cr. | 0.242 | 0.161 | $9 \%$ | 847 | 4.58 | 0.03 | 1.63 |
| Black R. | 0.531 | -1.483 | $45 \%$ | 23,532 | 4.88 | 0.36 | 45.25 |
| Genesee R. | 0.559 | -1.562 | $67 \%$ | 8,890 | 1.39 | 0.03 | 17.10 |
| Oswego R. | 0.467 | -1.614 | $67 \%$ | 15,244 | 1.15 | 0.02 | 29.31 |
| Salmon R. | 0.401 | -0.942 | $98 \%$ | 1,745 | 2.45 | 0.27 | 3.36 |
| dissolved mercury |  |  |  |  |  |  |  |
| 18-Mile Cr. | 0.595 | -1.046 | $26 \%$ | 344 | 1.86 | 0.01 | 0.66 |
| Black R. | 0.393 | -1.291 | $80 \%$ | 9,760 | 2.02 | 0.15 | 18.77 |
| Genesee R. | 0.346 | -1.378 | $40 \%$ | 2,262 | 0.35 | 0.01 | 4.35 |
| Oswego R. | 0.468 | -2.022 | $50 \%$ | 6,023 | 0.46 | 0.01 | 11.58 |
| Salmon R. | 0.048 | 0.053 | $1 \%$ | 1,452 | 2.04 | 0.22 | 2.79 |

Broken mercury thermometers have been seen in the Black River immediately downstream from the sampling point. An elevated sediment mercury sediment concentration ( $2.4 \mathrm{mg} / \mathrm{kg}$ ) was seen 5.2 km upstream from the sampling point on the Black River.

Concentrations of mercury in water may be contextualized through examining the distribution of mercury in sediments.


Figure 4. Mercury in Great Lakes Basin surficial sediments taken from the NYSDEC National Sediment Inventory.

## PCB Results

Measurement of PCBs was hampered by detection limits that were in most places high relative to the native concentrations. Table 7 shows all measured total PCB concentrations. Highlighted records have PCB concentrations that were less than 5 times the highest blank. Individual congener concentrations and detection levels are shown in the Appendix. Sample specific detection limits averaged by sites and homologs are also shown in the Appendix.

Concentration/discharge relations were calculated as $\log / \log$ linear equations and the slopes and intercepts were applied to the actual discharges that occurred over the period of record. Summary statistics were calculated and shown in Table 8. Yellow-shaded records are of low confidence due to inadequate detection limits and, for the Oswego and Black Rivers, poor correlations. It is likely that PCB concentrations in the Black, Genesee, Oswego, and Salmon Rivers are overestimates.

TABLE 7. PCB samples, results, and instantaneous loads.

| site | date | total, ng/L | disch., CFS | g/hr |
| :---: | :---: | :---: | :---: | :---: |
| 18-Mile Cr. | 5/1/07 | 103 | 210 | 1.92 |
| 18-Mile Cr. | $3 / 3 / 08$ | 109 | 101 | 0.52 |
| 18-Mile Cr. | $3 / 24 / 08$ | 33 | 273 | 0.83 |
| 18-Mile Cr. | $4 / 7 / 08$ | 56 | 301 | 1.59 |
| 18-Mile Cr. | $4 / 29 / 08$ | 93 | 70 | 1.03 |
| 18-Mile Cr. | $5 / 27 / 08$ | 145 | 107 | 1.32 |
| 18-Mile Cr. | $6 / 9 / 08$ | 383 | 96 | 3.15 |
| Black R. | $5 / 3 / 07$ | 1.8 | 8730 | 1.60 |
| Black R. | $10 / 24 / 07$ | 1.8 | 6980 | 1.30 |
| Black R. | $3 / 13 / 08$ | 0.77 | 12200 | 0.96 |
| Black R. | $5 / 28 / 08$ | 2.2 | 2660 | 0.61 |
| Black R. | $10 / 1 / 08$ | 1.8 | 1900 | 0.36 |
| Genesee R. | $5 / 2 / 07$ | 0.65 | 4250 | 0.28 |
| Genesee R. | $3 / 4 / 08$ | 0.37 | 4680 | 0.17 |
| Genesee R. | $4 / 8 / 08$ | 0.85 | 8030 | 0.69 |
| Genesee R. | $5 / 28 / 08$ | 2.3 | 1680 | 0.39 |
| Genesee R. | $6 / 10 / 08$ | 1.8 | 1417 | 0.26 |
| Genesee R. | $6 / 19 / 08$ | 1.5 | 827 | 0.13 |
| Genesee R. | $9 / 30 / 08$ | 0.84 | 1451 | 0.12 |
| Oswego R. | $5 / 2 / 07$ | 0.76 | 19700 | 1.53 |
| Oswego R. | $10 / 25 / 07$ | 0.57 | 9570 | 0.56 |
| Oswego R. | $1 / 10 / 08$ | 0.49 | 15200 | 0.76 |
| Oswego R. | $3 / 4 / 08$ | 0.15 | 11900 | 0.19 |
| Oswego R. | $3 / 25 / 08$ | 0.13 | 16600 | 0.22 |
| Oswego R. | $4 / 30 / 08$ | 1.9 | 11800 | 2.31 |
| Oswego R. | $9 / 30 / 08$ | 0.56 | 1160 | 0.07 |
| Salmon R. | $5 / 3 / 07$ | 0.30 | 1533 | 0.05 |
| Salmon R. | $1 / 10 / 08$ | 0.77 | 2581 | 0.20 |
| Salmon R. | $4 / 8 / 08$ | 0.45 | 2998 | 0.14 |
| Salmon R. | $5 / 1 / 08$ | 0.47 | 2096 | 0.10 |
| duplicate |  |  |  |  |
| Oswego R. | $4 / 30 / 08$ | 0.89 | 11800 | 1.07 |
| field blanks |  |  |  |  |
| Genesee R. | $9 / 30 / 08$ | 0.41 | 1451 |  |
| Oswego R. | $9 / 30 / 08$ | 0.38 | 1160 |  |
|  |  |  |  |  |
|  |  |  |  |  |

TABLE 8. PCB loads derived from log/log regressions.

|  | slope | intercept | r2 | total load (g) | g/sq km | g/capita | g/day |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 18-Mile Cr. | -0.75 | 3.58 | $53 \%$ | 15,720 | 85 | 0.52 | 30 |
| Black R. | -0.35 | 1.62 | $15 \%$ | 11,741 | 2.4 | 0.18 | 23 |
| Genesee R. | -0.70 | 2.35 | $66 \%$ | 2,631 | 0.41 | 0.008 | 5.1 |
| Oswego R. | -0.12 | 0.15 | $2 \%$ | 3,850 | 0.29 | 0.004 | 7.4 |
| Salmon R. | 0.92 | -3.40 | $69 \%$ | 333 | 0.47 | 0.051 | 0.64 |

Figure 5 shows PCBs in surficial sediments.


Figure 5. Total PCBs in Great Lakes Basin surficial sediments taken from the NYSDEC National Sediment Inventory.

## Pesticides Results

Table 9 shows the pesticides (and metabolites) that were measured here, pesticide groupings, and NYS water quality standards. Pesticide observations were spotty. Table 10 shows the strongest pesticide data - all observations at least five times greater than field blanks, and five times larger than method blanks. The most frequently observed pesticides were dieldrin, endosulfan, and the metabolite, heptachlor epoxide.

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Table 9. Pesticides, groupings, NYS Water Quality Standards

|  |  | WQS. |  | Wrouping | ngS. |
| :--- | :--- | :---: | :--- | :--- | :---: |
| pesticide | grouping | ng/L | pesticide | ng |  |
| BHC, alpha | BHC | 2 | $4,4^{\prime}-$ DDE | DDT | 0.007 |
| BHC, beta | BHC | 7 | 4,4 '-DDT | DDT | 0.01 |
| BHC, delta | BHC | 8 | Aldrin | Dieldrin | 1 |
| BHC, gamma | BHC | 8 | Dieldrin | Dieldrin | 0.0006 |
| Chlordane, cis | Chlordane | 0.02 | Endosulfan sulfate | Endosulfan | NA |
| Chlordane, trans | Chlordane | 0.02 | Endosulfan, alpha | Endosulfan | NA |
| Chlordane,oxy- | Chlordane | 0.02 | Endosulfan, beta | Endosulfan | NA |
| Nonachlor, cis- | Chlordane | 0.02 | Endrin | Endrin | 2 |
| Nonachlor, trans- | Chlordane | 0.02 | Endrin aldehyde | Endrin | NA |
| Hexachlorobenzene | Chlorobenzene | 0.03 | Endrin ketone | Endrin | NA |
| $2,4 '-D D D ~$ | DDT | NA | Heptachlor | Heptachlor | 0.2 |
| 2,4 '-DDE | DDT | NA | Heptachlor epoxide | Heptachlor | 0.3 |
| $2,4 '-D D T$ | DDT | NA | Methoxychlor | Methoxychlor | 30 |
| 4,4'-DDD | 0.08 | Mirex | Mirex | 0.001 |  |

Pesticide concentrations were highest in 18-Mile Creek, particularly on the 3/3/2008 sampling event. Most noteworthy was the high concentration of parent (4,4'-) DDT. DDT was banned in NYS in 1970, two years before being banned nationally by EPA.

TABLE 10. Pesticides concentrations (ng/L)

| Site | date | $\begin{aligned} & \text { 2,4'- } \\ & \text { DDD } \end{aligned}$ | $\begin{aligned} & \text { 2,4'- } \\ & \text { DDE } \end{aligned}$ | $\begin{aligned} & 2,4 '- \\ & \text { DDT } \end{aligned}$ | $\begin{aligned} & 4,4^{\prime} \\ & \text { DDD } \end{aligned}$ | $\begin{aligned} & 4,4^{\prime} \\ & \text { DDE } \end{aligned}$ | $\begin{aligned} & 4,4 \\ & \text { DDT } \end{aligned}$ | Aldrin | alpha <br> BHC | beta <br> BHC | delta <br> BHC | gamma <br> BHC |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 18-Mile Cr. | 5/1/07 | 0.076 |  |  | 0.156 | 0.342 |  | 0.008 |  |  |  |  |
|  | 3/3/08 | 4.19 | 0.500 | 10.800 | 6.22 | 39.2 | 71.8 |  |  |  | 0.009 | 0.075 |
|  | 3/24/08 | 0.086 |  | 0.068 | 0.163 | 1.21 | 0.435 |  |  |  |  | 0.076 |
|  | 4/7/08 |  |  |  | 0.223 | 0.550 | 0.189 |  | 0.138 | 0.185 | 0.026 | 0.142 |
|  | 4/29/08 |  |  |  | 0.179 | 0.234 |  |  |  | 0.059 | 0.014 | 0.094 |
|  | 5/27/08 |  |  |  | 0.232 | 0.233 |  |  | 0.041 | 0.053 |  | 0.078 |
|  | 6/9/08 |  |  |  | 0.406 | 0.456 | 0.524 | 0.246 |  |  |  |  |
| 18-Mile Cr. Avg Black R |  | 1.45 | 0.500 | 5.43 | 1.08 | 6.03 | 18.2 | 0.127 | 0.090 | 0.099 | 0.016 | 0.093 |
|  | 10/24/07 |  |  |  |  | 0.040 |  |  |  |  |  |  |
|  | 1/9/08 |  |  |  |  | 0.102 |  |  |  |  |  |  |
|  | 5/28/08 |  |  |  |  |  |  |  |  |  |  | 0.029 |
|  | 10/1/08 |  |  |  |  |  |  | 0.020 |  |  |  |  |
| Black R. Avg |  |  |  |  |  | 0.071 |  | 0.020 |  |  |  | 0.029 |
| Genesee R. | 5/2/07 | 0.024 |  |  |  | 0.154 |  |  |  |  |  |  |
|  | 3/4/08 |  |  | 0.025 | 0.036 | 0.234 | 0.119 |  |  |  |  |  |
|  | 4/8/08 |  |  |  |  | 0.203 | 0.115 |  | 0.038 |  |  | 0.030 |
|  | 6/10/08 |  |  |  |  | 0.055 |  |  |  |  |  |  |
|  | 6/19/08 |  |  |  |  | 0.097 |  |  |  |  |  |  |
| Genesee R. Avg |  | 0.024 |  | 0.025 | 0.036 | 0.149 | 0.117 |  | 0.038 |  |  | 0.030 |
| Oswego R. | 5/2/07 | 0.027 |  |  | 0.074 | 0.134 |  |  | 13.0 | 1.61 |  | 0.169 |
|  | 10/25/07 |  |  |  |  | 0.107 |  |  |  |  |  |  |
|  | 1/10/08 |  |  |  |  | 0.127 | 0.086 |  |  |  |  |  |
|  | 3/4/08 |  |  |  | 0.018 | 0.044 |  |  |  |  |  |  |
|  | 3/25/08 |  |  | 0.051 | 0.050 | 0.222 | 0.204 |  |  |  |  |  |
|  | 4/30/08 |  |  |  |  |  |  |  | 0.030 |  |  | 0.037 |
| Oswego R. Avg |  | 0.027 |  | 0.051 | 0.047 | 0.127 | 0.145 |  | 4.353 | 1.610 |  | 0.081 |
| Salmon R. | 1/10/08 |  |  |  |  | 0.041 |  |  |  |  |  |  |
| Salmon R. Avg |  |  |  |  |  | 0.041 |  |  |  |  |  |  |

Table 10. Continued.

| site | date | cis Chlordane | trans- <br> Chlordane | oxy- <br> Chlordane | Dieldrin | sulfate <br> Endosulfan | alpha, Endosulfan | beta, Endosulfan |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 18-Mile Cr. | 5/1/07 |  |  |  | 0.299 | 0.250 |  |  |
|  | 3/3/08 | 0.220 | 0.207 | 0.026 | 2.30 | 22.5 | 1.85 | 7.08 |
|  | 3/24/08 | 0.029 |  | 0.018 | 0.192 | 0.670 | 0.121 |  |
|  | 4/7/08 | 0.130 | 0.140 |  | 0.298 | 0.617 |  |  |
|  | 4/29/08 | 0.043 | 0.046 | 0.061 | 0.351 | 0.438 |  |  |
|  | 5/27/08 | 0.128 | 0.071 |  | 0.326 | 0.213 |  |  |
|  | 6/9/08 | 0.186 |  |  | 0.480 | 0.384 |  |  |
| 18-Mile Cr. Avg |  | 0.123 | 0.116 | 0.035 | 0.607 | 3.58 | 0.986 | 7.08 |
| Black R. | 10/24/07 |  |  |  | 0.028 | 0.187 |  |  |
|  | 1/9/08 |  | 0.052 |  | 0.031 | 0.115 |  |  |
|  | 5/28/08 |  |  |  | 0.023 | 0.079 |  |  |
| Black R. Avg |  |  | 0.052 |  | 0.027 | 0.127 |  |  |
| Genesee R. | 3/4/08 | 0.034 |  |  | 0.133 | 0.088 | 0.042 |  |
|  | 4/8/08 | 0.020 | 0.014 |  | 0.046 |  |  |  |
|  | 5/28/08 | 0.040 | 0.031 |  | 0.088 | 0.124 |  |  |
|  | 6/10/08 |  |  |  | 0.105 | 0.174 |  |  |
|  | 6/19/08 |  |  |  | 0.126 | 0.392 |  |  |
|  | 9/30/08 |  |  | 0.057 | 0.094 |  |  |  |
| Genesee R. Avg |  | 0.031 | 0.023 | 0.057 | 0.099 | 0.195 | 0.042 |  |
| Oswego R. | 5/2/07 |  |  |  | 0.092 |  |  |  |
|  | 10/25/07 |  | 0.016 |  | 0.051 | 0.173 |  |  |
|  | 1/10/08 |  | 0.016 |  | 0.070 | 0.195 |  |  |
|  | 3/4/08 | 0.015 |  |  | 0.059 | 0.125 | 0.051 |  |
|  | 3/25/08 |  |  |  | 0.075 | 0.180 | 0.078 |  |
|  | 4/30/08 | 0.015 | 0.014 |  | 0.093 | 0.148 |  |  |
| Oswego R. Avg |  | 0.015 | 0.015 |  | 0.076 | 0.161 | 0.065 |  |
| Salmon R. | 1/10/08 |  |  |  | 0.021 | 0.061 |  |  |
|  | 4/8/08 |  | 0.014 |  | 0.017 |  |  |  |
|  | 5/1/08 |  |  |  | 0.029 | 0.075 |  |  |
| Salmon R. Avg |  |  | 0.014 |  | 0.022 | 0.068 |  |  |

Table 10. Continued.

| site | date | Endrin | ketone, Endrin | Heptachlor | epoxide Heptachlor | Hexachlorobenzene | Methoxychlor | Mirex | cis+trans <br> Nonachlor |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 18-Mile Cr. | 5/1/07 |  |  |  | 0.029 |  |  |  |  |
|  | 3/3/08 | 0.333 | 0.282 | 0.042 | 0.052 | 0.189 | 0.045 | 0.036 | 0.243 |
|  | 3/24/08 |  | 0.028 | 0.011 | 0.033 |  |  |  |  |
|  | 4/7/08 |  | 0.031 | 0.043 | 0.034 |  |  |  | 0.051 |
|  | 4/29/08 |  | 0.019 |  | 0.046 |  |  |  | 0.041 |
|  | 5/27/08 |  |  |  | 0.042 |  |  | 0.018 | 0.091 |
|  | 6/9/08 |  |  | 0.095 | 0.056 |  |  | 0.072 |  |
| 18-Mile Cr. Avg |  | 0.333 | 0.090 | 0.048 | 0.042 | 0.189 | 0.045 | 0.042 | 0.117 |
| Black R. | 5/3/07 |  |  |  | 0.013 |  |  |  |  |
|  | 10/24/07 |  |  |  |  |  |  |  |  |
|  | 1/9/08 |  |  |  |  |  |  |  | 0.027 |
|  | 5/28/08 |  |  | 0.011 |  |  |  |  | 0.016 |
|  | 10/1/08 |  |  | 0.027 |  |  |  |  |  |
| Black R. Avg |  |  |  | 0.019 | 0.013 |  |  |  | 0.022 |
| Genesee R. | 5/2/07 |  |  |  | 0.024 | 0.786 |  |  |  |

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Table 11 shows the average ratio by which pesticides exceed the Water Quality Standard. Non-detections are censored. Cases where there were no observed instances of an exceedence are not shown. The 18-Mile Creek sampling site was directly adjacent to a busy and prosperous-looking farm.

Table 11. Exceedence ratio.

| pesticide | 18-Mile Cr. | Black R. | Genesee R. | Oswego R. | Salmon R. |
| :--- | :---: | :---: | :---: | :---: | :---: |
| BHC, alpha |  |  |  | 2.18 |  |
| Chlordane + Nonachlor | 14.4 | 2.38 | 3.93 | 1.33 | 1.65 |
| Hexachlorobenzene | 6.3 |  | 26.2 |  |  |
| 4,4'-DDE | 13.5 |  |  |  |  |
| 4,4'-DDD | 862 | 10.1 | 21.2 | 18.1 | 5.86 |
| 4,4'-DDT | 1824 |  | 11.7 | 14.5 |  |
| Dieldrin | 1011 | 45.6 | 164 | 127 | 37.2 |
| Mirex | 42 |  |  | 138 |  |

Table 12. Instantaneous loads for pesticides with Water Quality Standards (g/day). Sites with maximum average loads are highlighted.


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| 3/4/08 | 0.412 | 2.68 | 1.36 |  |  |  |  | 0.389 | 1.52 |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| site date | $\begin{aligned} & 4,4^{\prime} \\ & \text { DDD } \end{aligned}$ | $\begin{aligned} & 4,4 \\ & \text { DDE } \end{aligned}$ | $\begin{aligned} & \hline 4,4 \\ & \text { DDT } \end{aligned}$ | Aldrin | alpha <br> BHC | beta <br> BHC | $\begin{aligned} & \mathrm{D}+\mathrm{G} \\ & \mathrm{BHC} \end{aligned}$ | Chlordane | Dieldrin | Endrin |
| 4/8/08 |  | 3.99 | 2.26 |  | 0.747 |  | 0.589 | 0.668 | 0.904 |  |
| 5/28/08 |  |  |  |  |  |  |  | 0.292 | 0.362 |  |
| 6/10/08 |  | 0.191 |  |  |  |  |  |  | 0.364 |  |
| 6/19/08 |  | 0.196 |  |  |  |  |  |  | 0.255 |  |
| 9/30/08 |  |  |  |  |  |  |  | 0.202 | 0.334 |  |
| Genesee R. Avg | 0.412 | 1.73 | 1.81 |  | 0.747 |  | 0.589 | 0.388 | 0.624 |  |
| Oswego R. 5/2/07 | 3.57 | 6.46 |  |  | 627 | 77.6 | 8.146 |  | 4.43 |  |
| 10/25/07 |  | 2.51 |  |  |  |  |  | 0.375 | 1.19 |  |
| 1/10/08 |  | 4.72 | 3.20 |  |  |  |  | 0.595 | 2.60 | 0.669 |
| 3/4/08 | 0.524 | 1.28 |  |  |  |  |  | 0.437 | 1.72 |  |
| 3/25/08 | 2.03 | 9.02 | 8.29 |  |  |  |  |  | 3.05 |  |
| 4/30/08 |  |  |  |  | 1.73 |  | 2.108 | 1.67 | 5.34 |  |
| Oswego R. Avg | 2.04 | 4.80 | 5.74 |  | 314 | 77.6 | 5.127 | 0.770 | 3.06 | 0.669 |
| Salmon R. 1/10/08 |  | 0.259 |  |  |  |  |  |  | 0.133 |  |
| 4/8/08 |  |  |  |  |  |  |  | 0.103 | 0.125 |  |
| 5/1/08 |  |  |  |  |  |  |  |  | 0.149 |  |
| Salmon R. Avg |  | 0.259 |  |  |  |  |  | 0.103 | 0.135 |  |

Table 12. Continued.

| site | date | Heptachlor | epoxide Heptachlor | Hexachlorobenzene | Methoxychlor | Mirex | cis+trans Nonachlor |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 18-Mile Cr. | 5/1/07 |  | 0.013 |  |  |  |  |
|  | 3/3/08 | 0.005 | 0.006 | 0.022 | 0.005 | 0.004 | 0.028 |
|  | 3/24/08 | 0.007 | 0.020 |  |  |  |  |
|  | 4/7/08 | 0.029 | 0.023 |  |  |  | 0.035 |
|  | 4/29/08 |  | 0.012 |  |  |  | 0.011 |
|  | 5/27/08 |  | 0.009 |  |  | 0.004 | 0.020 |
|  | 6/9/08 | 0.019 | 0.011 |  |  | 0.014 |  |
| 18-Mile Cr. Avg |  | 0.015 | 0.013 | 0.022 | 0.005 | 0.007 | 0.023 |
| Black R. | 5/3/07 |  | 0.278 |  |  |  |  |
|  | 1/9/08 |  |  |  |  |  | 0.925 |
|  | 5/28/08 | 0.072 |  |  |  |  | 0.104 |
|  | 10/1/08 | 0.126 |  |  |  |  |  |
| Black R. Avg |  | 0.099 | 0.278 |  |  |  | 0.515 |
| Genesee R. | 5/2/07 |  | 0.249 | 8.17 |  |  |  |
|  | 3/4/08 |  | 0.481 |  |  |  |  |
|  | 4/8/08 | 0.393 | 0.334 |  |  |  | 0.963 |
|  | 5/28/08 |  | 0.115 |  |  |  | 0.284 |
|  | 6/10/08 |  | 0.111 |  |  |  |  |
|  | 6/19/08 |  | 0.065 |  |  |  |  |
| Genesee R. Avg |  | 0.393 | 0.226 | 8.17 |  |  | 0.623 |
| Oswego R. | 5/2/07 | 0.251 | 0.945 |  |  |  | 0.675 |
|  | 10/25/07 |  |  |  |  | 6.18 | 0.375 |
|  | 1/10/08 |  | 0.855 |  |  | 0.446 | 0.409 |
|  | 3/4/08 |  | 0.379 |  |  |  |  |
|  | 3/25/08 |  | 0.406 |  |  |  |  |
|  | 4/30/08 |  | 0.924 |  |  |  | 0.375 |
| Oswego R. Avg Salmon R. |  | 0.251 | 0.702 |  |  | 3.31 | 0.458 |
|  | 4/8/08 |  |  |  |  |  | 0.286 |
|  | 5/1/08 |  |  |  |  |  | 0.067 |

Sediment data (taken from the NYSDEC National Sediment Inventory) are shown for 4,4'-DDT, dieldrin, and total chlordanes in Figures 6, 7, and 8.


Figure 6. 4,4'-DDT in Great Lakes Basin surficial sediments.


Figure 7. Dieldin in Great Lakes Basin surficial sediments.


Figure 8. Total chlordanes in Great Lakes Basin surficial sediments.

## PCDD/F Results

"Dioxins" is shorthand for polychlorinated dibenzo-p-dioxins (PCDDs) and polychlorinated dibenzofurans (PCDFs). There are theoretically 75 distinct PCDDs (called congeners) and 135 PCDF congeners. There are an equal number of polybrominated analogs. Potentially, there are 337 mixed chlorinated and brominated dioxins (PXDDs) and 647 furans (PXDFs). In the regulatory world, however, only 7 PCDDs and 10 PCDFs are routinely measured. These regulatory target PCDD/Fs have chlorine atoms attached to $2,3,7$, and 8 positions on the dibenzo-p-dioxin or dibenzofuran skeleton. Up to four more chlorine atoms can be attached to other positions.

It is not the intention of this report to discuss dioxin health effects but we will point out that some of the congeners are extraordinarily toxic.[4-11] New York State's Water Quality Standard for the sum ${ }^{\text {a }}$ of all $17 \mathrm{PCDD} /$ Fs is 0.6 femtograms/L. ${ }^{\text {b }}$ Table 13 lists the 17 regulatory PCDD/Fs along with their toxic equivalency factors (TEFs) from the World Health Organization [12] and the bioaccumulation equivalency factors (BEFs) used by NYS. From time to time an expert panel considers new informations and amends the TEFs as appropriate. The effects of these changes vary with the composition of the sample but on average they have reduced the overall calculated toxicity by about $19 \%$.

[^0]Almost all the change was between WHO94 and 1998 revision. The mean relative percent difference between all of this project's total TEQs caluculated with WHO98 and WHO05 is $4.45 \%$. Table 13 also gives a Congener Order for each congener that will be used subsequently in place of the longer congener name.

Individual congener concentrations and detection limits are shown in the Appendix.
Table 13. Dioxins, names, congener orders, evolving TEF, and BEFs.

| congener <br> order | congener | WHO94 | WHO98 | WHO05 | BEF |
| ---: | :--- | ---: | ---: | ---: | ---: |
| 1 | 2,3,7,8-TCDD | 1 | 1 | 1 | 1 |
| 2 | $1,2,3,7,8-P e C D D$ | 0.5 | 1 | 1 | 0.9 |
| 3 | $1,2,3,4,7,8-H x C D D$ | 0.1 | 0.1 | 0.1 | 0.3 |
| 4 | $1,2,3,6,7,8-H x C D D$ | 0.1 | 0.1 | 0.1 | 0.1 |
| 5 | $1,2,3,7,8,9-H x C D D$ | 0.1 | 0.1 | 0.1 | 0.1 |
| 6 | $1,2,3,4,6,7,8-H p C D D$ | 0.01 | 0.01 | 0.01 | 0.05 |
| 7 | OCDD | 0.001 | 0.0001 | 0.0003 | 0.01 |
| 8 | $2,3,7,8-$ TCDF | 0.1 | 0.1 | 0.1 | 0.8 |
| 9 | $1,2,3,7,8-\mathrm{PeCDF}$ | 0.05 | 0.05 | 0.03 | 0.2 |
| 10 | $2,3,4,7,8-\mathrm{PeCDF}$ | 0.5 | 0.5 | 0.3 | 1.6 |
| 11 | $1,2,3,4,7,8-H x C D F$ | 0.1 | 0.1 | 0.1 | 0.08 |
| 12 | $1,2,3,6,7,8-H x C D F$ | 0.1 | 0.1 | 0.1 | 0.2 |
| 13 | $2,3,4,6,7,8-H x C D F$ | 0.1 | 0.1 | 0.1 | 0.7 |
| 14 | $1,2,3,7,8,9-H x C D F$ | 0.1 | 0.1 | 0.1 | 0.6 |
| 15 | $1,2,3,4,6,7,8-H p C D F$ | 0.01 | 0.01 | 0.01 | 0.01 |
| 16 | $1,2,3,4,7,8,9-H p C D F$ | 0.01 | 0.01 | 0.01 | 0.4 |
| 17 | OCDF | 0.001 | 0.0001 | 0.0003 | 0.02 |

Table 14. Sites, dates, times, and $L$ processed for dioxin samples.

| site | date | start | end | L filtered |
| :--- | ---: | ---: | ---: | ---: |
| 18-Mile Cr. | $5 / 1 / 07$ | $12: 47$ | $14: 16$ | 336 |
| 18-Mile Cr. | $3 / 3 / 08$ | $14: 36$ | $15: 29$ | 284 |
| 18-Mile Cr. | $3 / 24 / 08$ | $13: 19$ | $16: 59$ | 923 |
| 18-Mile Cr. | $4 / 7 / 08$ | $14: 02$ | $17: 30$ | 1,169 |
| 18-Mile Cr. | $4 / 29 / 08$ | $13: 21$ | $17: 00$ | 1,244 |
| 18-Mile Cr. | $5 / 27 / 08$ | $13: 58$ | $15: 48$ | 543 |
| 18-Mile Cr. DU | $5 / 27 / 08$ | $13: 58$ | $15: 57$ | 591 |
| 18-Mile Cr. | $6 / 9 / 08$ | $14: 06$ | $15: 32$ | 430 |
| Black R. | $5 / 3 / 07$ | $13: 55$ | $16: 30$ | 647 |
| Black R. | $10 / 24 / 07$ | $11: 45$ | $14: 30$ | 305 |
| Black R. | $1 / 9 / 08$ | $14: 06$ | $15: 27$ | 157 |
| Black R. | $3 / 13 / 08$ | $11: 08$ | $13: 38$ | 594 |
| Black R. | $5 / 28 / 08$ | $14: 33$ | $18: 51$ | 1,279 |
| Black R. | $6 / 10 / 08$ | $15: 10$ | $17: 20$ | 759 |
| Equip. Blank | $2 / 27 / 08$ | $10: 27$ |  | 1 |
| Equip. Blank | $6 / 11 / 08$ | $11: 07$ |  | 1 |
| Genesee R. | $5 / 2 / 07$ | $7: 49$ | $9: 20$ | 309 |
| Genesee R. | $3 / 4 / 08$ | $8: 44$ | $10: 49$ | 518 |
| Genesee R. | $4 / 8 / 08$ | $8: 20$ | $8: 49$ | 151 |
| Genesee R. | $5 / 28 / 08$ | $7: 59$ | $10: 10$ | 759 |
| Genesee R. | $6 / 10 / 08$ | $8: 40$ | $10: 20$ | 604 |
| Genesee R. | $6 / 19 / 08$ | $14: 33$ | $16: 09$ | 580 |
| Genesee R. | $9 / 30 / 08$ | $11: 40$ | $12: 38$ | 395 |


| Oswego R. | $5 / 2 / 07$ | $14: 06$ | $16: 35$ | 599 |
| :--- | ---: | ---: | :--- | ---: |
| site | date | start | end | L filtered |
| Oswego R. | $10 / 25 / 07$ | $10: 18$ | $14: 53$ | 830 |
| Oswego R. | $1 / 10 / 08$ | $8: 18$ | $11: 00$ | 630 |
| Oswego R. | $3 / 4 / 08$ | $14: 29$ | $16: 41$ | 562 |
| Oswego R. | $3 / 25 / 08$ | $14: 25$ | $18: 03$ | 921 |
| Oswego R. | $4 / 30 / 08$ | $12: 09$ | $16: 51$ | 1,411 |
| Oswego R. | $9 / 30 / 08$ | $16: 36$ | $18: 22$ | 679 |
| Salmon R. | $5 / 3 / 07$ | $7: 58$ | $10: 43$ | 652 |
| Salmon R. | $1 / 10 / 08$ | $13: 30$ | $16: 36$ | 654 |
| Salmon R. | $4 / 8 / 08$ | $13: 32$ | $15: 07$ | 515 |
| Salmon R. | $5 / 1 / 08$ | $8: 43$ | $14: 25$ | 1,726 |

Table 15 shows for sample the total PCDD/F concentration calculated with each of the three TEFs and the bioaccumulation equivalency factor that is used in the NYSDEC Water Quality Standard. The later WHO TEFs reflect evolving science but the WHO 94 value is in regulation.

Table 15. Total TEQs in $\mathrm{fg} / \mathrm{L}$ calculated with three different TEFs.

| site | date | WHO 94, <br> BEF | WHO 98, <br> BEF | WHO 05, <br> BEF |
| :--- | ---: | :---: | :---: | :---: |
| 18-Mile Cr. | $5 / 1 / 07$ | 84.7 | 91.2 | 68.6 |
| 18-Mile Cr. | $3 / 3 / 08$ | 527 | 576 | 436 |
| 18-Mile Cr. | $3 / 24 / 08$ | 64.3 | 69.2 | 50.9 |
| 18-Mile Cr. | 4/7/08 | 22.3 | 26.9 | 27.0 |
| 18-Mile Cr. | $4 / 29 / 08$ | 65.0 | 68.2 | 48.6 |
| 18-Mile Cr. | $5 / 27 / 08$ | 326 | 341 | 246 |
| 18-Mile Cr. | $6 / 9 / / 08$ | 510 | 535 | 382 |
| Black R. | $5 / 3 / 07$ | 7.21 | 8.44 | 6.86 |
| Black R. | $10 / 24 / 07$ | 31.4 | 37.5 | 29.7 |
| Black R. | $1 / 9 / 08$ | 63.7 | 63.4 | 46.6 |
| Black R. | $3 / 13 / 08$ | 2.29 | 3.41 | 3.41 |
| Black R. | $5 / 28 / 08$ | 7.57 | 8.68 | 6.49 |
| Black R. | $6 / 10 / 08$ | 9.24 | 9.20 | 6.14 |
| Genesee R. | $5 / 2 / 07$ | 36.1 | 41.7 | 34.6 |
| Genesee R. | $3 / 4 / 08$ | 29.3 | 35.2 | 30.6 |
| Genesee R. | $4 / 8 / 08$ | 40.3 | 46.8 | 35.6 |
| Genesee R. | $5 / 28 / 08$ | 89.5 | 99.9 | 75.5 |
| Genesee R. | $6 / 10 / 08$ | 65.1 | 72.3 | 56.4 |
| Genesee R. | $6 / 19 / 08$ | 68.2 | 76.8 | 58.7 |
| Genesee R. | $9 / 30 / 08$ | 67.9 | 73.7 | 55.4 |
| Oswego R. | $5 / 2 / 07$ | 27.1 | 31.4 | 24.8 |
| Oswego R. | $10 / 25 / 07$ | 37.3 | 43.8 | 34.7 |
| Oswego R. | $1 / 10 / 08$ | 15.9 | 19.0 | 15.0 |
| Oswego R. | $3 / 4 / 08$ | 0.81 | 0.79 | 0.78 |
| Oswego R. | $3 / 25 / 08$ | 13.2 | 16.2 | 13.3 |
| Oswego R. | $4 / 30 / 08$ | 23.3 | 27.1 | 22.0 |
| Oswego R. | $9 / 30 / 08$ | 13.7 | 15.6 | 12.3 |
| Salmon R. | $5 / 3 / 07$ | 9.30 | 11.2 | 8.63 |
| Salmon R. | $1 / 10 / 08$ | 12.4 | 12.4 | 8.35 |
| Salmon R. | $4 / 8 / 08$ | 18.1 | 21.3 | 16.3 |
| Salmon R. | $5 / 1 / 08$ | 7.91 | 9.29 | 7.04 |
|  |  |  |  |  |

Table 16 shows TEQs for PCDD/F samples. The table gives:
TEQ, pg recov. Sum of the products of PCDD/F congeners and the WHO98 TEF. ratio

POC, mg/L
TEQ, fg/L
trapping efficiency
Ratio of the $\Sigma$ TEQs calculated where non-detections are assigned values of 0 or the sample specific detection limit. As the ratio gets lower, the possible error from the non-detections increases.
Concentration of particulate organic carbon in unfiltered water.
TEQ concentration in $\mathrm{fg} / \mathrm{L}$.
The efficiency of the cartridge filter in trapping particles is assessed by $1-(\mathrm{F} / \mathrm{R})$ where $\mathrm{F}=$ the POC concentration in water that has passed through the cartridge filter and R is the raw unfiltered water.
corrected TEQ, fg/L TEQ concentration divided by trapping efficiency. This value is used in load calculations.

Appendix A shows concentrations for each PCDD/F congener and PCDD/F homolog.
Table 16. Dioxins in NYS Tributaries to Lake Ontario, total TEQ (WHO98). Highlighted records have TEQ pg recoveries less than 5 x the maximum field blank.

| site | date | TEQ <br> pg recov. | ratio | POC <br> mg/L | TEQ, <br> fg/L | trapping <br> efficiency | corrected <br> TEQ, fg/L |
| :--- | ---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 18-Mile Cr. | $5 / 1 / 07$ | 74 | $97 \%$ | 0.20 | 220 | $52 \%$ | 424 |
| 18-Mile Cr. | $3 / 3 / 08$ | 353 | $90 \%$ | 1.70 | 1,246 | $49 \%$ | 2,556 |
| 18-Mile Cr. | $3 / 24 / 08$ | 137 | $85 \%$ | 0.21 | 149 | $49 \%$ | 305 |
| 18-Mile Cr. | $4 / 7 / 08$ | 181 | $66 \%$ | 0.33 | 155 | $65 \%$ | 238 |
| 18-Mile Cr. | $4 / 29 / 08$ | 155 | $84 \%$ | 0.22 | 125 | $41 \%$ | 304 |
| 18-Mile Cr. | $5 / 27 / 08$ | 433 | $82 \%$ | 0.44 | 760 | $37 \%$ | 2,066 |
| 18-Mile Cr. | $6 / 9 / 08$ | 497 | $87 \%$ | 0.60 | 1,156 | $73 \%$ | 1,588 |
| Black R. | $5 / 3 / 07$ | 9 | $83 \%$ | 0.37 | 14 | $79 \%$ | 18 |
| Black R. | $10 / 24 / 07$ | 24 | $94 \%$ |  | 80 | $65 \%$ | 123 |
| Black R. | $1 / 9 / 08$ | 20 | $75 \%$ | 1.13 | 126 | $81 \%$ | 155 |
| Black R. | $3 / 13 / 08$ | 4 | $45 \%$ | 0.16 | 7 | $52 \%$ | 13 |
| Black R. | $5 / 28 / 08$ | 16 | $74 \%$ | 0.19 | 13 | $56 \%$ | 23 |
| Black R. | $6 / 10 / 08$ | 12 | $56 \%$ | 0.30 | 16 | $57 \%$ | 28 |
| Genesee R. | $5 / 2 / 07$ | 21 | $93 \%$ | 0.32 | 69 | $62 \%$ | 110 |
| Genesee R. | $3 / 4 / 08$ | 32 | $88 \%$ | 0.23 | 61 | $62 \%$ | 98 |
| Genesee R. | $4 / 8 / 08$ | 11 | $69 \%$ | 0.80 | 73 | $50 \%$ | 146 |
| Genesee R. | $5 / 28 / 08$ | 147 | $85 \%$ | 0.4 | 193 | $69 \%$ | 280 |
| Genesee R. | $6 / 10 / 08$ | 81 | $92 \%$ | 0.40 | 134 | $69 \%$ | 194 |
| Genesee R. | $6 / 19 / 08$ | 85 | $87 \%$ | 0.50 | 147 | $68 \%$ | 214 |
| Genesee R. | $9 / 30 / 08$ | 43 | $91 \%$ | 0.59 | 108 | $62 \%$ | 174 |
| Oswego R. | $5 / 2 / 07$ | 32 | $88 \%$ | 0.19 | 53 | $63 \%$ | 83 |
| Oswego R. | $10 / 25 / 07$ | 81 | $98 \%$ | 0.43 | 98 | $73 \%$ | 133 |
| Oswego R. | $1 / 10 / 08$ | 20 | $89 \%$ | 0.25 | 33 | $67 \%$ | 49 |
| Oswego R. | $3 / 4 / 08$ | 3 | $33 \%$ | 0.12 | 6 | $51 \%$ | 12 |
| Oswego R. | $3 / 25 / 08$ | 23 | $81 \%$ | 0.25 | 25 | $41 \%$ | 61 |
| Oswego R. | $4 / 30 / 08$ | 78 | $83 \%$ | 0.27 | 55 | $64 \%$ | 86 |
| Oswego R. | $9 / 30 / 08$ | 20 | $86 \%$ |  | 29 | $60 \%$ | 49 |
| Salmon R. | $5 / 3 / 07$ | 10 | $85 \%$ | 0.12 | 15 | $75 \%$ | 20 |
| Salmon R. | $1 / 10 / 08$ | 13 | $69 \%$ | 0.25 | 20 | $81 \%$ | 25 |
| Salmon R. | $4 / 8 / 08$ | 15 | $65 \%$ | 0.29 | 29 | $82 \%$ | 36 |
| Salmon R. | $5 / 1 / 08$ | 21 | $69 \%$ | 0.12 | 12 | $63 \%$ | 19 |
|  |  |  |  | 23 |  |  |  |
|  |  |  |  |  |  |  |  |

Table 17 shows POC corrected regressions for total TEQ. Yellow highlights signify low confidence due to poor correlations in the concentration/discharge relationship.

Table 17. Dioxin loads derived from log/log regressions.

|  |  |  | total load |  |  | ug/sq |  |
| :--- | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| Total TEQ | slope | intercept | R2 | (ug) | km | ug/capita | TEQ |
| 18-Mile Cr. | -1.02 | 5.03 | 33 | 127,480 | 689 | 4.19 | 246 |
| Black R. | 0.42 | -0.04 | 7.3 | 225,798 | 47 | 3.46 | 436 |
| Genesee R. | -0.32 | 3.31 | 52 | 469,277 | 73 | 1.41 | 906 |
| Oswego R. | 0.06 | 1.52 | 0.5 | 460,851 | 35 | 0.46 | 890 |
| Salmon R. | 0.78 | -1.23 | 68 | 17,949 | 25 | 2.73 | 34.7 |

The loading rate ( $\mu \mathrm{g} / \mathrm{sq} \mathrm{km}$ of watershed) was very much greater in $18-\mathrm{Mile} \mathrm{Cr}$. than elsewhere but the total loadings were greatest in the larger rivers. Other than as laboratory standards or for research purposes, dioxins were never intentionally manufactured. There are, however, a large variety of natural [13], industrial [14, 15], and inadvertent or accidental [16-18] events that have been shown to generate dioxins.

Figure 9 shows an overview of PCDD/F concentrations (as WHO 98 TEQs) from the NYSDEC National Sediment Inventory database of surficial sediment samples taken in the Lake Erie/Lake Ontario/St. Lawrence drainage where at least six of the 17 congeners were quantified. The size of the circles indicates concentration. Relatively high concentrations occur in the Niagara River, 18-Mile Creek, Onondaga Lake, and across Lake Ontario.

Congener patterns may provide insights into $\mathrm{PCDD} / \mathrm{F}$ sources. Congener concentrations (normalized by WHO 98 TEFs) were ranked. Each sample was labeled with the congener order (see Table 13) contributing the highest and second highest TEQ. Table 18 shows the number of instances where various congeners were the largest contributor to total TEQ. Congeners 1 and 2 both have TEFs of 1 giving them both the highest toxicities. However, congener 6 , while only having a TEF of 0.01 , is much more abundant and is most often the largest single contributor of total TEQ.

Processes differ in the patterns of congeners formed. For example, the manufacture of trichlorophenoxy acetic acid used as a component in the Vietnam War era defoliant Agent Orange was particularly effective at producing 2,3,7,8-TCDD (Congener 1).[9, 16] Uncontrolled fires are often rich sources congener 10.[15, 19] Bleaching kraft process paper pulp with elemental chlorine resulted in formation of congeners 1 and 8.[20] While the PCDD/Fs are lumped together under the TEQ rubric, they are actually chemically distinct substances with different sources.


Figure 9. Total PCDD/Fs in Great Lakes Basin surficial sediments. NYSDEC National Sediment Inventory.

Table 18. Instances where particular PCDD/F congeners are the greatest contributor to total TEQ (WHO 98) in the NY portion of the Great Lakes Basin.

| Congener <br> Order | instances |
| :---: | :---: |
| 6 | 138 |
| 1 | 97 |
| 10 | 90 |
| 8 | 85 |
| 11 | 64 |
| 2 | 52 |
| 15 | 3 |
| 4 | 1 |
| 12 | 1 |
| 13 | 1 |

Forty-two patterns of first/second congener dominance were found. Of the 532 sediment samples, $87 \%$ had one of 18 couplet patterns (Table 19). The first entry in Table 19 (couplet 8-10) shows that congener 8 was the largest and congener 10 was the second largest contributor to total TEQ.

| Congener <br> couplets | instances |
| :--- | :---: |
| $8-10$ | 73 |
| $10-8$ | 48 |
| $1-2$ | 41 |
| $6-4$ | 39 |
| $1-11$ | 32 |
| $2-1$ | 25 |
| $6-10$ | 25 |
| $6-2$ | 22 |
| $11-10$ | 21 |
| $10-2$ | 20 |
| $6-11$ | 20 |
| $6-15$ | 20 |
| $1-10$ | 18 |
| $11-15$ | 14 |
| $11-6$ | 12 |
| $2-6$ | 12 |
| $2-10$ | 11 |
| $10-6$ | 10 |

Table 19. Instances where particular PCDD/F congener couplets are the greatest contributor to total TEQ (WHO 98) in the NY portion of the Great Lakes Basin.


Figure 10. PCDD/F sediment concentrations and congener patterns (in TEQ) from the Niagara River and vicinity. NYSDEC National Sediment Inventory.

The Occidental Durez site in North Tonawanda manufactured phenolic resins using the Raschig-Hooker Process. Benzene is chlorinated by exposing it to a hot air- HCl mix. The resulting chlorobenzenes were then exposed to NaOH to form phenol, NaCl and water.[21] Side reactions occurred whereby PCDD/Fs and octachlorostyrene were also formed. Durez discharged liquid waste to sewers which emptied into the Niagara River at the Pettit Flume, in North Tonawanda. A sediment core taken from 740 m downstream
at Fisherman's Park in 1999 shows the highest sediment PCDD/F levels measured in NYS. The Pettit Flume was remediated between 1989 and 1995.


Figure 11. PCDD/F concentrations from a sediment core at Fisherman's Park. Surficial concentrations are much lower than deeper ones. NYSDEC National Sediment Inventory.

When the relative abundances of the congeners are plotted, however, the three samples appear virtually identical. Dominant congeners here were 11 and 15 . A number of samples dominated by congeners 1 and 11 clusters around Cayuga Island and the Love Canal area. Sediments from throughout Lake Ontario show the 1-11 pattern. Sediments downstream from Pettit Flume in the Caygua Little River (Niagara Falls) also show congener 11 dominance. Congener 1 is dominant in Caygua Creek proper.


Figure 12. PCDD/F fingerprints from the Fisherman's Park sediment core. NYSDEC National Sediment Inventory. Legend indicates core depth in m. See Figure 11.

Tonawanda Creek naturally ran into the Niagara River but the lower portion of the creek has been converted into the Western end of the Erie Canal. During navigation season water flows from the Niagara into the lower Tonawanda Creek and through the canal. At Lockport (and Gasport) three gates let a total of 65 cfs from the Erie Canal into 18-Mile Creek (and the East Branch of 18-Mile Creek).

## 18-Mile Creek - Sediment



Figure 13. Surficial PCDD/F sediment concentrations and dominant congener couplets in the Erie Canal and in 18-Mile Creek. NYSDEC National Sediment Inventory.

PCDD/F concentrations in surficial Tonawanda Creek sediments are dominated by congeners 2 and 1 . Samples taken within the City of Lockport show several new congeners, particularly 6 and 11. PCDD/F patterns change in Lockport suggesting sources with unusually high relative amounts of congeners 11 and 15 . Congener 6 was also very abundant, particularly in the lower strata of a sediment core.

## 18-Mile Creek - Water

Table 20 shows the rank orders of the top five PCDD/F congeners from the 18 -Mile Creek water (suspended sediment) samples.

Table 20. Ranks of congener contributions to total TEQ, 18-Mile Creek. The highest rank is 1.

| Congener order | 1 | 2 | 4 | 5 | 6 | 10 | 11 | 12 | 15 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| 18-Mile Cr.-3/24/2008-SA |  |  | 5 |  | 1 | 2 | 3 |  | 4 |
| 18-Mile Cr.-5/1/2007-SA |  |  | 5 |  | 1 | 2 | 3 |  | 4 |
| 18-Mile Cr.-6/9/2008-SA |  |  | 5 |  | 1 | 2 | 3 |  | 4 |
| 18-Mile Cr.-3/3/2008-SA |  |  |  |  | 1 | 2 | 4 | 3 | 5 |
| 18-Mile Cr.-5/27/2008-DU |  |  | 5 |  | 1 | 2 | 4 |  | 3 |
| 18-Mile Cr.-4/7/2008-SA |  | 5 | 4 |  | 1 |  | 2 |  | 3 |
| 18-Mile Cr.-4/29/2008-SA |  |  | 5 |  | 2 | 1 | 3 |  | 4 |
| 18-Mile Cr.-5/27/2008-SA |  |  | 5 |  | 2 | 1 | 4 |  | 3 |

## Genesee River - Sediment

Comparatively few sediment samples were taken upstream from the sampling site on the Genesee River. These are shown in Figure 14.


Figure 14. Sediment samples and dominant congener couplets in the vicinity of the Genesee River sampling site off the Andrews St. Bridge. NYSDEC National Sediment Inventory.

## Genesee River - Water

Samples taken during canal navigation season had higher PCDD/F concentrations than those taken in the winter ( 158 vs $68 \mathrm{fg} / \mathrm{L}$ TEQ). Congener abundances were also different. Congeners 11 and 15 were more abundant during navigation season. Congener 2 was more prevalent in non-navigation season.

Table 21. Ranks of congener contributions to total TEQ, Genesee River. The highest rank is 1.

|  | 1 | 2 | 4 | 5 | 6 | 10 | 11 | 12 | 15 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Genesee R.-5/28/2008-SA |  | 4 |  |  | 2 | 1 | 3 |  | 5 |
| Genesee R.-6/10/2008-SA |  | 4 |  |  | 2 | 1 | 3 |  | 5 |
| Genesee R.-6/19/2008-SA |  | 4 |  |  | 2 | 1 | 3 |  | 5 |
| Genesee R.-9/30/2008-SA |  | 4 | 5 |  | 2 | 1 | 3 |  |  |
| Genesee R.-4/8/2008-SA |  | 2 | 5 |  | 3 | 1 | 4 |  |  |
| Genesee R.-5/2/2007-SA | 4 | 1 |  |  | 2 | 3 | 5 |  |  |
| Genesee R.-3/4/2008-SA | 4 | 1 |  | 5 | 2 | 3 |  |  |  |

## Oswego River - Sediment



Figure 15. Surficial and core sediment data from the Oswego River and vicinity. NYSDEC National Sediment Inventory.

Sediment cores from this area demonstrate the range of congener patters stemming from different sources. Figure 16 shows a core in Lake Ontario. This is the same pattern that appeared in the Cayuga Island/Love Canal samples indicated in Figure 10.


Figure 16. Sediment core in Lake Ontario off from Oswego River. NYSDEC National Sediment Inventory.


Figure 17. Sediment core in Onondaga Lake near waste water treatment plant. NYSDEC National Sediment Inventory.

Dominance of congener 10 is associated with combustion and incineration activities.


Figure 18. Site of sediment core in Onondaga Lake near wastewater treatment plant. NYSDEC National Sediment Inventory.

Most of the sediment samples from the southern portion of Onondaga Lake were rich in congeners 10 and 8 . This pattern could have been generated by incineration.


Figure 19. Sediment core off Battle Island in the Oswego River upstream of water sampling site at Lock 8. ng/kg. NYSDEC National Sediment Inventory.

## Oswego River - Water

The dioxin abundances observed at Lock 8 from suspended sediment (water) samples included congeners 6 and 10 seen in upstream sediments. Congener 2 was relatively more abundant in the suspended sediment samples than it had been in upstream bottom sediments..

Table 22. Ranks of congener contributions to total TEQ, Oswego River. The highest rank is 1.

|  | 1 | 2 | 4 | 5 | 6 | 10 | 11 | 12 | 15 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Oswego R.-10/25/2007-SA |  | 2 | 4 | 5 | 1 | 3 |  |  |  |
| Oswego R.-4/30/2008-SA |  | 2 | 4 |  | 1 | 3 |  |  | 5 |
| Oswego R.-5/2/2007-SA |  | 3 | 4 | 5 | 1 | 2 |  |  |  |
| Oswego R.-9/30/2008-SA |  | 3 | 4 | 5 | 1 | 2 |  |  |  |
| Oswego R.-1/10/2008-SA | 1 | 5 | 4 | 3 | 2 |  |  |  |  |
| Oswego R.-3/25/2008-SA | 1 | 5 | 4 | 3 | 2 |  |  |  |  |



Figure 20. PCDD/F relative abundances, Oswego River at Oswego

## Salmon River - Water

There are no sediment data for dioxins from the Salmon River. This project put significant effort into sampling the Salmon R. but the quality of the results was generally poor.

Table 23. Comparisons of level of effort and success in PCDD/F sampling in the five tributaries.

| Site | Avg L <br> sampled | Avg ratio, <br> ND=0/ND=DL |
| :--- | :---: | :---: |
| 18 Mile Cr., Corwin | 690 | 0.84 |
| Black R., Watertown | 624 | 0.71 |
| Genesee R., Rochester | 487 | 0.86 |
| Oswego R., Oswego | 826 | 0.79 |
| Salmon River, Pulaski | 887 | 0.72 |

Table 24. Ranks of congener contributions to total TEQ, Salmon River. The highest rank is 1.

|  | 1 | 2 | 4 | 5 | 6 | 10 | 11 | 12 | 15 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Salmon R.-4/8/2008-SA |  | 2 |  | 5 | 3 | 1 | 4 |  |  |
| Salmon R.-5/1/2008-SA |  | 2 |  | 5 | 3 | 1 | 4 |  |  |
| Salmon R.-1/10/2008-SA |  |  | 5 | 4 | 3 | 1 | 2 |  |  |
| Salmon R.-5/3/2007-SA |  | 1 | 5 | 4 | 3 | 2 |  |  |  |

Congener 10 often appears abundant from combustion or incineration activities. For example, it was the most abundant congener in the ash and dust from 9/11/01 World Trade Center disaster. These samples were taken from a relatively pristine area where much of the local economy revolves around sport fishing. It is conceivable that wood boilers, wood stoves, and back-yard burn barrels may be contributing to this PCDD/F signal.[15]


Figure 21. Characteristic patterns of PCDD/F congener abundances from backyard burn barrels and from wood stoves.

## Black River - Sediment

Figure 22 shows the locations of the Black River (and the Salmon River) sampling sites. Dioxin concentrations are low in Black River sediments and sediment patterns are dominated by congeners 6,15 , and 2 .


Figure 22. Sediment and water sampling sites on the Salmon and Black Rivers. NYSDEC National Sediment Inventory.

## Black River - Water

Table 25 shows the ranks of congeners found in Black River water samples. While congener 15 was important in some of the sediment samples, it was a minor component of the suspended sediment PCDD/Fs.

Table 25. Ranks of congener contributions to total TEQ, Black River. The highest rank is 1.

|  | 1 | 2 | 4 | 5 | 6 | 10 | 11 | 12 | 15 |
| :--- | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: | ---: |
| Black R.-1/9/2008-SA |  |  | 3 | 4 | 1 | 2 | 5 |  |  |
| Black R.-10/24/2007-SA |  | 2 | 4 | 5 | 1 | 3 |  |  |  |
| Black R.-5/3/2007-SA |  | 1 | 4 | 5 | 2 | 3 |  |  |  |
| Black R.-5/28/2008-SA |  | 2 | 4 | 5 | 3 | 1 |  |  |  |



Figure 23. PCDD/F relative abundances, Black River at Watertown.

## Does Congener 6 Come From Dechlorination of Octachlorodioxin?

We have seen frequent examples of the importance of heptachlorodioxin (congener 6) to total TEQ. What are the sources of congener 6? EPA surveyed the literature and published characteristic PCDD/F congener patterns from a variety of industrial activities ranging from crematoria to municipal solid waste (MSW) incinerators to copper smelters [14, 15]. Of 53 different source patterns, only one (diesel trucks) showed congener 6 making a significant contribution to total TEQ. Another pattern ("Baltimore tunnel") also represents diesel trucks but here congener 6 was not even the $5^{\text {th }}$ most important congener.

Table 26. Congener couplets identified in EPA's survey of PCDD/F sources.

| Source | Congener couplets |
| :--- | :--- |
| Kraft pulp sludge | $1-8$ |
| Kraft pulp | $1-8$ |
| auto exhaust, unleaded | $1-10$ |
| Pb smelter, after scrubber | $1-10$ |
| tire combustion | $1-10$ |
| chlor-alkaki, DOW | $1-11$ |
| ferrous foundries | $2-1$ |
| residential wood stove chimney soot, Canadian | $2-4$ |
| cars, leaded | $2-10$ |
| MSW bottom ash | $2-10$ |
| utility, oil | $2-10$ |
| bleached-kraft mill sludge in wood residue boilers | $8-10$ |
| landfill flare | $8-10$ |
| Baltimore tunnel | $10-1$ |


| Source | Congener couplets |
| :---: | :---: |
| auto exhause, diesel | 10-1 |
| Pb smelter, before scrubber | 10-1 |
| Al smelter, 2 | 10-2 |
| Al smelter, 5 | 10-2 |
| Al smelter, 6 | 10-2 |
| cement kiln | 10-2 |
| cement kiln, haz waste, high temp | 10-2 |
| crematorium | 10-2 |
| forest fires | 10-2 |
| indust. wood burner | 10-2 |
| indust. wood burner, ash | 10-2 |
| large municipal waste combustors | 10-2 |
| MSW fly ash | 10-2 |
| oil fired industrial boilers | 10-2 |
| residential coal combustors | 10-2 |
| Al smelter, 1 | 10-8 |
| cement kiln, 2000 | 10-8 |
| cement kiln, haz waste, low temp | 10-8 |
| cement kiln, non haz waste | 10-8 |
| lightweight aggregate kiln | 10-8 |
| sewage sludge incineration | 10-8 |
| utility, coal | 10-8 |
| Al smelter, 3 | 10-11 |
| Al smelter, 4 | 10-11 |
| burn barrel | 10-11 |
| chlor-alkaki, PPG | 10-11 |
| Cu smelter, Chemetco | 10-11 |
| halogen acid furnace, 2000 | 10-11 |
| Haz waste incinerators, 1993-6 | 10-11 |
| Hot Sided ESP boilers, 1993-6 | 10-11 |
| Hot Sided ESP boilers, 2000 | 10-11 |
| burn barrel, avid recycler | 10-13 |
| burn barrel, avid recycler, non recycler | 10-13 |
| industrial wood combustors | 10-14 |
| Cu smelter, Franklin | 11-10 |
| hazardous waste incinerators, 2000 | 11-10 |
| metal recovery facility ash/soil, open burn sites | 14-10 |
| metal recovery facility fly ash | 14-10 |
| diesel truck | 14-6 |

If octachlorodioxin loses a single chlorine, the two possibilities are $1,2,3,4,6,7,9-\mathrm{HpCDD}$ or $1,2,3,4,6,7,8-\mathrm{HpCDD}$ (congener 6 ). $1,2,3,4,6,7,9-\mathrm{HpCDD}$ is not normally reported under regulatory sampling because it lacks the $2,3,7,8$ - required chlorination positions. Monodechlorination of the most abundant congener would produce congener 6. Sediment data (see Figure 19) suggests that the proportion of congener 6 is greater in deeper and older sediments. It is possible that a former technology produced significantly more congener 6 that those EPA surveyed or it may be that monodechlorination of octachlorodioxin is occurring. Bacterial dechlorination of refractory organics, like octachlorodioxin, usually does not occur until the target substance is very abundant and constitutes an important energy source. But what are the alternative explanations?

## Appendix - PCB and PCDD/F Data

Results are shown for $\operatorname{PCB}(\mathrm{ng} / \mathrm{L})$ and $\mathrm{PCDD} / \mathrm{F}(\mathrm{fg} / \mathrm{L})$ congeners where there were detections. PCB congeners are identified by the standard Ball-Schmitter Zell (BZ) numbers. Samples for PCBs are designated by code to save space.

Appendix Table 1. PCB sample code

| sample | code | sample | code |
| :---: | :---: | :---: | :---: |
| 18-Mile Cr.-3/24/2008-SA | A-1 | Oswego R.-1/10/2008-SA | D-1 |
| 18-Mile Cr.-3/3/2008-SA | A-2 | Oswego R.-10/25/2007-SA | D-2 |
| 18-Mile Cr.-4/29/2008-SA | A-3 | Oswego R.-3/25/2008-SA | D-3 |
| 18-Mile Cr.-4/7/2008-SA | A-4 | Oswego R.-3/4/2008-SA | D-4 |
| 18-Mile Cr.-5/1/2007-SA | A-5 | Oswego R.-4/30/2008-DU | D-5 |
| 18-Mile Cr.-5/27/2008-SA | A-6 | Oswego R.-4/30/2008-SA | D-6 |
| Black R.-10/1/2008-SA | B-1 | Oswego R.-5/2/2007-SA | D-7 |
| Black R.-10/24/2007-SA | B-2 | Oswego R.-9/30/2008-FB | D-8 |
| Black R.-3/13/2008-SA | B-3 | Oswego R.-9/30/2008-SA | D-9 |
| Black R.-5/28/2008-SA | B-4 | Salmon R.-1/10/2008-SA | E-1 |
| Black R.-5/3/2007-SA | B-5 | Salmon R.-4/8/2008-SA | E-2 |
| Genesee R.-3/4/2008-SA | C-1 | Salmon R.-5/1/2008-SA | E-3 |
| Genesee R.-4/8/2008-SA | C-2 | Salmon R.-5/3/2007-SA | E-4 |
| Genesee R.-5/2/2007-SA | C-3 |  |  |
| Genesee R.-5/28/2008-SA | C-4 |  |  |
| Genesee R.-6/10/2008-SA | C-5 |  |  |
| Genesee R.-9/30/2008-FB | C-6 |  |  |
| Genesee R.-9/30/2008-SA | C-7 |  |  |

Appendix Table 2. PCB detection limits, site and homolog averages, pg/L

| homolog | 18-Mile <br> Cr. | Black <br> R. | Genesee <br> R. | Oswego <br> R. | Salmon <br> R. |
| :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 30.7 | 11.2 | 4.14 | 9.29 | 3.32 |
| 2 | 22.9 | 34 | 26.2 | 33.7 | 26.7 |
| 3 | 7.7 | 10.3 | 7.12 | 12.1 | 6.34 |
| 4 | 7.66 | 10.2 | 5.69 | 8.89 | 5.96 |
| 5 | 6.82 | 7.26 | 4.7 | 7.46 | 5.16 |
| 6 | 7.36 | 12.2 | 6.01 | 10.6 | 9.09 |
| 7 | 4.79 | 8.08 | 6.41 | 10 | 7.73 |
| 8 | 7.59 | 11.6 | 6.25 | 11.2 | 5.91 |
| 9 | 6.67 | 11.2 | 7.55 | 12.4 | 6.26 |
| 10 | 6.38 | 14.2 | 8.94 | 20.2 | 6.26 |

Appendix Table 3. PCBs, ng/L

| BZ | A-1 | A-2 | A-3 | A-4 | A-5 | A-6 | B-1 | B-2 | B-3 | B-4 | B-5 | C-1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.102 | 0.179 |  | 0.089 |  | 0.621 | 0.0513 | 0.142 | 0.0564 | 0.839 | 0.412 |  |
| 2 |  |  |  |  |  |  | 0.0023 | 0.00244 |  | 0.0088 |  |  |
| 3 | 0.0103 |  |  |  |  | 0.0583 |  | 0.0179 | 0.0071 | 0.119 |  |  |
| 4 | 0.409 | 2.84 |  | 3.8 | 0.118 | 9.69 | 0.32 | 5.66 | 0.268 | 11.2 | 10.2 |  |
| 5 |  |  |  | 0.371 |  |  |  |  |  |  |  |  |
| 6 |  | 0.257 |  | 0.227 |  | 1.12 |  | 0.349 | 0.023 | 1.44 | 0.952 |  |
| 7 |  |  |  |  |  | 0.097 |  | 0.043 |  | 0.105 | 0.092 |  |
| 8 |  | 0.277 |  |  |  | 1.08 | 0.087 | 0.34 | 0.065 | 1.61 | 0.933 |  |
| 9 |  |  |  |  |  | 0.147 |  | 0.054 |  | 0.16 | 0.124 |  |
| 10 |  |  |  | 0.064 |  | 0.277 |  | 0.141 | 0.018 | 0.196 | 0.302 |  |
| 11 |  |  |  | 0.06 |  | 0.124 | 0.035 |  |  | 0.128 | 0.142 |  |
| 12 |  |  |  |  |  | 0.595 |  | 0.281 |  | 0.713 | 0.819 |  |
| 15 |  | 0.444 |  | 1.39 |  | 1.77 | 0.03 | 0.935 |  | 2.1 | 2.48 |  |
| 16 |  | 0.089 |  | 0.287 |  | 0.599 | 0.0211 | 0.309 |  | 0.591 | 0.79 |  |
| 17 |  | 0.953 |  | 2.66 | 0.026 | 4.8 | 0.0775 | 2.58 | 0.0399 | 5.54 | 6.71 |  |
| 18 | 0.076 | 0.691 | 0.048 | 2.04 | 0.038 | 3.88 | 0.0948 | 2.04 | 0.0503 | 4.26 | 5.14 |  |
| 19 | 0.099 | 0.796 | 0.059 | 1.66 | 0.043 | 3.43 | 0.0986 | 1.93 | 0.0576 | 3.74 | 4.05 |  |
| 20 | 0.0721 | 0.538 | 0.095 | 2.6 | 0.0318 | 3.2 | 0.0896 | 1.75 | 0.0552 | 3.49 | 4.85 | 0.0145 |
| 21 | 0.0165 | 0.04 |  | 0.115 |  | 0.201 | 0.0249 | 0.102 | 0.0142 | 0.181 | 0.279 | 0.0056 |
| 22 | 0.0175 | 0.095 | 0.015 | 0.328 |  | 0.524 | 0.0256 | 0.238 | 0.0155 | 0.545 | 0.719 |  |
| 24 |  |  |  |  |  |  |  |  |  |  |  |  |
| 25 | 0.0147 | 0.308 | 0.015 | 1.54 |  | 2.1 | 0.0187 | 1.28 | 0.0105 | 2.75 | 3.15 |  |
| 26 | 0.0415 | 0.528 | 0.037 | 2.54 | 0.0098 | 4.15 | 0.0445 | 2.24 | 0.0254 | 4.66 | 5.97 |  |
| 27 | 0.051 | 0.335 | 0.037 | 1.02 |  | 2.19 | 0.0471 | 1.11 | 0.0246 | 2.16 | 2.86 |  |
| 31 | 0.0828 | 0.721 | 0.084 | 3.15 | 0.0246 | 4.47 | 0.0944 | 2.44 | 0.0632 | 5.33 | 6.52 | 0.0104 |
| 32 | 0.033 | 0.418 |  | 1.28 |  | 2.43 | 0.0307 | 1.32 | 0.0236 | 2.95 | 3.51 |  |
| 34 |  |  |  | 0.039 |  | 0.0705 |  | 0.0387 |  | 0.0681 | 0.102 |  |
| 35 |  |  |  | 0.017 |  |  |  |  |  | 0.0115 | 0.0179 |  |
| 37 |  | 0.061 |  | 0.338 | 0.011 | 0.216 |  | 0.129 | 0.0102 | 0.235 | 0.39 |  |
| 39 |  |  |  | 0.028 |  |  |  | 0.0131 |  |  | 0.0374 |  |
| 40 |  | 0.791 |  | 3.34 |  | 2.39 | 0.0382 | 1.46 | 0.0257 | 2.53 | 4 |  |
| 42 |  | 0.429 |  | 1.93 |  | 1.35 |  | 0.804 | 0.0125 | 1.38 | 2.41 |  |
| 43 |  | 0.055 |  | 0.243 |  | 0.235 | 0.0029 | 0.144 |  | 0.19 | 0.383 |  |
| 44 | 0.0742 | 1.63 | 0.153 | 7.18 | 0.053 | 5.43 | 0.0891 | 3.22 | 0.0511 | 5.65 | 9.39 | 0.028 |
| 45 |  | 0.282 |  | 1.29 |  | 1.5 | 0.0267 | 0.871 | 0.0127 | 1.56 | 2.37 |  |
| 46 |  | 0.079 |  | 0.372 |  | 0.464 | 0.0055 | 0.266 |  | 0.444 | 0.715 |  |
| 48 |  | 0.141 |  | 0.561 |  | 0.453 | 0.0117 | 0.265 |  | 0.398 | 0.741 |  |
| 49 | 0.0503 | 1.31 | 0.057 | 5.68 | 0.034 | 4.25 | 0.0607 | 2.58 | 0.0405 | 4.46 | 7.31 |  |
| 50 | 0.0366 | 0.25 | 0.032 | 1.17 |  | 1.88 | 0.0351 | 1.03 | 0.0202 | 1.95 | 2.61 |  |
| 52 | 0.114 | 1.92 | 0.127 | 7.92 | 0.057 | 6.29 | 0.108 | 3.82 | 0.0773 | 6.82 | 10.1 |  |
| 54 |  | 0.025 |  | 0.082 |  |  |  | 0.0526 |  | 0.0938 | 0.128 |  |
| 56 | 0.0139 | 0.304 | 0.03 | 1.36 |  | 0.558 | 0.0175 | 0.452 | 0.0152 | 0.731 | 1.19 |  |
| 57 |  |  |  | 0.078 |  | 0.0421 |  |  |  | 0.042 |  |  |
| 59 |  | 0.118 |  | 0.513 |  | 0.422 |  | 0.233 |  | 0.352 | 0.711 |  |
| 60 | 0.0083 |  | 0.02 | 0.597 |  | 0.294 | 0.0117 | 0.2 | 0.0079 | 0.277 | 0.557 |  |
| 61 |  | 1.38 | 0.157 | 5.7 | 0.058 | 2.86 | 0.0822 | 1.98 | 0.0702 | 3.24 | 5.42 | 0.022 |
| 63 |  | 0.059 |  | 0.246 |  | 0.133 |  | 0.082 |  | 0.128 | 0.248 |  |
| 64 | 0.0325 | 0.618 | 0.051 | 2.71 | 0.022 | 1.86 | 0.0381 | 1.12 | 0.0266 | 1.9 | 2.89 |  |
| 66 |  | 0.76 | 0.073 | 3.31 | 0.029 | 1.5 | 0.0402 | 1.03 | 0.0333 | 1.71 | 2.97 | 0.0135 |
| 67 |  |  |  | 0.113 |  | 0.0547 |  |  |  | 0.052 | 0.11 |  |
| 68 |  |  |  | 0.1 |  | 0.0539 | 0.0034 | 0.0358 |  | 0.051 | 0.101 |  |

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Appendix Table 3. Continued

| BZ | A-1 | A-2 | A-3 | A-4 | A-5 | A-6 | B-1 | B-2 | B-3 | B-4 | B-5 | C-1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 72 |  | 0.029 |  | 0.125 |  | 0.0709 |  | 0.0449 |  | 0.067 | 0.129 |  |
| 77 |  | 0.07 |  | 0.293 |  | 0.096 |  | 0.0792 |  | 0.105 | 0.22 |  |
| 79 |  |  |  |  |  | 0.0168 |  |  |  |  |  |  |
| 81 |  |  |  |  |  | 0.0021 |  | 0.00215 |  |  |  |  |
| 82 |  | 0.168 |  | 0.532 |  | 0.148 |  | 0.159 | 0.0043 | 0.186 | 0.378 |  |
| 83 | 0.0193 | 0.729 | 0.054 | 2.45 |  | 0.955 | 0.0255 | 0.733 | 0.0198 | 0.907 | 2.08 |  |
| 84 |  | 0.315 | 0.02 | 1.31 |  | 0.695 | 0.0136 | 0.452 |  | 0.667 | 1.36 |  |
| 85 |  | 0.239 | 0.126 | 0.769 |  | 0.257 |  | 0.207 | 0.0083 | 0.287 | 0.637 |  |
| 86 | 0.0315 | 0.715 | 0.063 | 2.38 |  | 0.934 | 0.0293 | 0.712 | 0.0294 | 0.926 | 2.02 |  |
| 88 |  | 0.22 |  | 0.904 |  | 0.424 | 0.0091 | 0.285 | 0.0064 | 0.416 | 0.874 |  |
| 89 |  | 0.025 |  | 0.097 |  | 0.0505 |  | 0.0318 |  | 0.0429 | 0.0933 |  |
| 90 | 0.0455 | 0.861 | 0.083 | 2.99 | 0.025 | 1.26 | 0.0436 | 0.917 | 0.0526 | 1.31 | 2.62 | 0.021 |
| 92 |  | 0.205 |  | 0.744 |  | 0.34 | 0.0089 | 0.232 |  | 0.325 | 0.711 |  |
| 93 |  | 0.114 |  | 0.421 |  | 0.23 |  | 0.149 |  | 0.205 | 0.443 |  |
| 94 |  |  |  | 0.087 |  |  |  | 0.0284 |  | 0.0419 | 0.0851 |  |
| 95 | 0.0475 | 0.672 | 0.058 | 2.74 | 0.02 | 1.55 | 0.0388 | 0.969 | 0.0387 | 1.55 | 2.92 | 0.015 |
| 96 |  | 0.019 |  |  |  | 0.0476 |  | 0.0269 |  | 0.0379 | 0.0907 |  |
| 103 |  |  |  | 0.059 |  | 0.0316 |  | 0.0206 |  | 0.0273 | 0.0598 |  |
| 104 |  |  |  |  |  |  |  | 0.00283 |  | 0.0041 |  |  |
| 105 | 0.0128 | 0.502 | 0.029 | 1.37 |  | 0.351 | 0.0135 | 0.347 | 0.0133 | 0.412 | 0.917 |  |
| 107 |  | 0.074 |  | 0.22 |  | 0.0645 | 0.00174 | 0.057 |  | 0.0647 | 0.155 |  |
| 108 |  | 0.033 |  | 0.097 |  | 0.0298 |  | 0.0259 |  | 0.0301 | 0.068 |  |
| 110 | 0.0501 | 1.38 |  | 4.78 | 0.027 | 1.95 | 0.0502 | 1.6 | 0.0482 | 2.05 | 4.1 | 0.0222 |
| 114 |  | 0.039 |  | 0.086 |  |  |  | 0.0236 |  | 0.024 | 0.0623 |  |
| 118 | 0.0332 | 0.995 | 0.065 | 2.84 | 0.023 | 0.804 | 0.0277 | 0.772 | 0.0344 | 0.971 | 1.99 | 0.016 |
| 120 |  |  |  |  |  |  |  |  |  |  | 0.00549 |  |
| 122 |  | 0.021 |  | 0.056 |  | 0.0166 |  | 0.0158 |  | 0.012 |  |  |
| 123 |  | 0.021 |  | 0.06 |  | 0.0186 | 0.00118 | 0.0193 |  | 0.0176 | 0.0439 |  |
| 126 |  |  |  |  |  |  |  | 0.00136 |  |  | 0.00422 |  |
| 128 |  | 0.119 |  | 0.288 |  | 0.0693 |  | 0.0675 |  | 0.0779 | 0.19 |  |
| 129 | 0.0665 | 0.874 | 0.083 | 1.92 | 0.034 | 0.446 | 0.0397 | 0.461 | 0.0534 | 0.489 | 1.18 | 0.049 |
| 130 |  | 0.048 |  | 0.113 |  |  |  | 0.0285 |  | 0.03 | 0.073 |  |
| 131 |  |  |  |  |  | 0.0072 |  | 0.0072 |  |  |  |  |
| 132 |  | 0.294 |  | 0.693 |  | 0.184 |  | 0.182 | 0.0196 | 0.21 | 0.47 |  |
| 133 |  |  |  | 0.036 |  | 0.0103 |  | 0.0088 |  | 0.0104 |  |  |
| 134 |  | 0.042 |  | 0.114 |  | 0.0317 |  | 0.0303 |  | 0.0337 | 0.0714 |  |
| 135 | 0.0271 | 0.256 |  | 0.522 |  | 0.172 | 0.0136 | 0.135 | 0.0209 | 0.186 | 0.395 |  |
| 136 |  | 0.082 |  | 0.186 |  | 0.0706 | 0.0043 | 0.0549 |  | 0.0685 | 0.158 |  |
| 137 |  | 0.044 |  | 0.106 |  |  |  | 0.025 |  | 0.0215 | 0.0556 |  |
| 139 |  |  |  | 0.038 |  |  |  | 0.0097 |  |  |  |  |
| 141 |  | 0.156 |  | 0.26 |  | 0.0665 | 0.0061 | 0.0627 |  | 0.0716 | 0.159 |  |
| 144 |  |  |  | 0.057 |  | 0.0175 | 0.0014 | 0.0145 |  | 0.0184 | 0.0391 |  |
| 146 |  | 0.098 |  | 0.215 |  | 0.0575 | 0.0052 | 0.0544 | 0.0084 | 0.0617 | 0.149 |  |
| 147 | 0.0487 | 0.603 | 0.051 | 1.37 | 0.025 | 0.387 | 0.0276 | 0.336 | 0.0446 | 0.426 | 0.933 | 0.038 |
| 148 |  |  |  |  |  | 0.00272 |  | 0.00207 |  | 0.0026 |  |  |
| 150 |  |  |  |  |  |  |  |  |  | 0.0018 |  |  |
| 152 |  |  |  |  |  |  |  | 0.00127 |  | 0.0019 |  |  |
| 153 | 0.0528 | 0.639 | 0.056 | 1.28 | 0.033 | 0.316 | 0.0301 | 0.302 | 0.0508 | 0.325 | 0.765 | 0.045 |
| 154 |  |  |  | 0.033 |  | 0.0103 |  | 0.00855 |  |  | 0.0243 |  |
| 156 |  | 0.093 |  | 0.241 |  | 0.0427 |  | 0.0477 |  | 0.0461 | 0.122 |  |

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Appendix Table 3. Continued

| BZ | A-1 | A-2 | A-3 | A-4 | A-5 | A-6 | B-1 | B-2 | B-3 | B-4 | B-5 | C-1 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 158 |  | 0.082 |  | 0.165 |  | 0.0407 | 0.0029 | 0.0407 |  | 0.0419 | 0.104 |  |
| 159 |  |  |  |  |  | 0.00223 |  | 0.00211 |  |  |  |  |
| 164 |  | 0.044 |  | 0.098 |  | 0.0261 |  | 0.0247 |  | 0.0268 | 0.066 |  |
| 167 |  | 0.028 |  | 0.0656 |  | 0.0136 |  | 0.0143 |  | 0.0154 | 0.0335 |  |
| 170 |  | 0.171 |  | 0.45 |  | 0.0565 | 0.011 | 0.0511 | 0.0138 | 0.0665 | 0.143 |  |
| 171 |  | 0.052 |  | 0.115 |  |  | 0.00286 | 0.0155 |  | 0.0191 | 0.0446 |  |
| 172 |  | 0.028 |  | 0.075 |  |  | 0.00176 | 0.0096 |  | 0.0123 | 0.0284 |  |
| 174 |  | 0.2 |  | 0.398 |  | 0.0584 | 0.01 | 0.0587 | 0.0163 | 0.0669 | 0.149 |  |
| 175 |  |  |  |  |  |  |  | 0.00224 |  |  | 0.00629 |  |
| 176 |  | 0.028 |  | 0.053 |  | 0.00849 |  | 0.00721 |  |  | 0.0211 |  |
| 177 |  | 0.104 |  | 0.237 |  | 0.0386 | 0.00557 | 0.0363 | 0.0082 | 0.0417 | 0.0969 |  |
| 178 |  | 0.043 |  | 0.093 |  | 0.0188 |  | 0.0176 | 0.0035 | 0.019 | 0.0456 |  |
| 179 |  | 0.083 |  | 0.161 |  | 0.0329 |  | 0.0294 | 0.0073 | 0.0352 | 0.0812 |  |
| 180 | 0.0287 | 0.371 | 0.03 | 1.06 | 0.026 | 0.135 | 0.0267 | 0.131 | 0.0346 | 0.143 | 0.357 | 0.04 |
| 183 |  | 0.106 |  | 0.244 |  | 0.0376 | 0.00597 | 0.0292 | 0.0093 | 0.0418 | 0.0968 |  |
| 187 | 0.0154 | 0.273 | 0.019 | 0.642 |  | 0.103 | 0.0152 | 0.0954 | 0.023 | 0.119 | 0.264 | 0.025 |
| 189 |  |  |  | 0.0196 |  |  |  | 0.00206 |  | 0.0033 | 0.00484 |  |
| 190 |  | 0.032 |  | 0.085 |  | 0.0113 |  | 0.00982 |  | 0.0112 | 0.0267 |  |
| 191 |  |  |  | 0.019 |  |  |  | 0.00185 |  |  | 0.00535 |  |
| 194 |  | 0.068 |  | 0.387 |  | 0.0302 | 0.00461 | 0.0333 | 0.0049 | 0.0372 | 0.0895 |  |
| 195 |  | 0.028 |  | 0.128 |  | 0.0107 |  | 0.0122 |  | 0.0135 | 0.0318 |  |
| 196 |  | 0.046 |  | 0.195 |  |  |  | 0.017 |  | 0.0195 | 0.0445 |  |
| 198 |  | 0.1 |  | 0.434 |  | 0.0465 |  | 0.0453 |  | 0.054 | 0.128 |  |
| 200 |  |  |  | 0.051 |  |  |  |  |  | 0.0071 |  |  |
| 201 |  |  |  | 0.05 |  | 0.00472 |  |  |  | 0.0069 | 0.0136 |  |
| 202 |  |  |  | 0.081 |  | 0.0102 |  | 0.0103 |  | 0.0112 | 0.0287 |  |
| 203 |  | 0.069 |  | 0.307 |  | 0.0257 | 0.00368 | 0.0281 |  | 0.0323 | 0.0795 |  |
| 205 |  |  |  | 0.018 |  |  |  | 0.00146 |  |  | 0.00358 |  |
| 206 |  | 0.036 |  | 0.272 |  | 0.0222 |  | 0.0274 |  | 0.039 | 0.0845 |  |
| 207 |  |  |  | 0.025 |  | 0.00301 |  | 0.003 |  |  | 0.00977 |  |
| 208 |  |  |  | 0.094 |  |  |  |  |  | 0.0158 | 0.0335 |  |
| 209 |  | 0.03 |  | 0.262 |  | 0.0247 |  | 0.0321 |  | 0.0468 | 0.119 |  |
| Total | 1.83 | 32.5 | 1.83 | 109 | 0.768 | 92.9 | 2.25 | 55.7 | 1.79 | 103 | 145 | 0.365 |

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Appendix Table 3. Continued

| BZ | C-2 | C-3 | C-4 | C-5 | C-6 | C-7 | D-1 | D-2 | D-3 | D-4 | D-5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 0.0029 | 0.0052 | 0.0085 |  |  |  |  |  |  | 0.011 | 0.00317 |
| 2 |  |  | 0.0042 |  |  |  |  |  |  |  |  |
| 3 | 0.0032 | 0.0032 |  |  |  |  |  |  |  |  |  |
| 4 |  |  | 0.111 | 0.163 |  | 0.163 |  |  |  |  | 0.0316 |
| 8 | 0.022 |  | 0.03 | 0.021 |  |  |  |  |  |  |  |
| 11 | 0.026 |  | 0.055 | 0.0261 |  |  |  |  |  |  |  |
| 15 |  |  | 0.045 | 0.048 |  |  |  |  |  |  |  |
| 16 |  |  |  |  |  |  |  |  |  |  | 0.0146 |
| 17 | 0.0138 | 0.0113 | 0.0585 | 0.0638 |  |  |  |  |  |  | 0.0191 |
| 18 | 0.0245 | 0.0154 | 0.0664 | 0.0792 |  |  |  |  |  |  | 0.0317 |
| 19 |  | 0.0062 | 0.0404 | 0.0552 |  |  |  |  |  |  | 0.0077 |
| 20 | 0.0388 | 0.0227 | 0.0844 | 0.069 | 0.0141 | 0.077 | 0.0228 | 0.0219 | 0.0165 | 0.0081 | 0.0456 |
| 21 | 0.0109 | 0.0084 | 0.0226 | 0.0139 |  |  |  |  |  |  | 0.0178 |
| 22 | 0.00506 | 0.0072 | 0.021 | 0.0135 |  |  |  |  |  |  | 0.0134 |
| 24 |  |  |  | 0.0141 |  |  |  |  |  |  |  |
| 25 | 0.00185 |  | 0.0174 | 0.0232 |  |  |  |  |  |  | 0.0085 |
| 26 | 0.00685 | 0.0071 | 0.0344 | 0.0399 |  |  |  |  |  |  | 0.015 |
| 27 |  |  |  | 0.0247 |  |  |  |  |  |  |  |
| 31 | 0.0283 | 0.0197 | 0.068 | 0.0643 | 0.0107 | 0.068 | 0.019 | 0.0159 | 0.0139 | 0.0071 | 0.0369 |
| 32 |  | 0.007 | 0.0323 |  |  |  |  |  |  |  |  |
| 37 | 0.00361 | 0.0068 | 0.0225 | 0.0139 |  |  |  |  |  |  | 0.0124 |
| 40 | 0.01088 | 0.0114 | 0.0543 | 0.0477 |  |  |  |  |  |  |  |
| 42 |  | 0.0095 |  |  |  |  |  |  |  |  | 0.0111 |
| 44 | 0.037 |  | 0.116 | 0.104 |  | 0.0947 | 0.061 | 0.079 | 0.025 | 0.021 | 0.0466 |
| 45 | 0.005 |  | 0.0263 | 0.0266 |  |  |  |  |  |  |  |
| 46 |  |  | 0.0085 | 0.0078 |  |  |  |  |  |  | 0.0023 |
| 48 |  |  | 0.0104 | 0.0109 |  |  |  |  |  |  | 0.0053 |
| 49 | 0.01523 | 0.0156 | 0.0751 | 0.0769 |  | 0.0467 |  |  |  |  | 0.0263 |
| 50 | 0.00163 | 0.0047 | 0.025 |  |  |  |  |  |  |  | 0.0065 |
| 52 | 0.0131 | 0.0321 | 0.124 | 0.12 | 0.0145 |  | 0.034 | 0.033 |  | 0.017 | 0.0454 |
| 56 | 0.00248 | 0.0073 | 0.028 | 0.0176 |  |  |  |  |  |  | 0.0103 |
| 59 | 0.00063 |  |  | 0.0079 |  |  |  |  |  |  |  |
| 60 | 0.00152 |  | 0.016 | 0.0073 |  |  |  |  |  |  | 0.0055 |
| 61 | 0.0319 | 0.0281 | 0.095 | 0.0619 |  |  | 0.0294 | 0.0294 | 0.021 |  | 0.0413 |
| 64 | 0.01096 | 0.0106 | 0.042 | 0.041 |  |  |  |  |  |  | 0.0144 |
| 66 | 0.00564 | 0.0155 | 0.065 | 0.0397 |  |  | 0.0135 | 0.0152 |  |  | 0.0198 |
| 68 | 0.00199 |  |  |  |  |  |  |  |  |  | 0.003 |
| 77 |  |  |  | 0.0048 |  |  |  |  |  |  | 0.0033 |
| 82 |  |  |  | 0.0081 |  |  |  |  |  |  | 0.0035 |
| 83 | 0.00661 | 0.0151 | 0.0495 | 0.0378 |  |  |  |  |  |  | 0.0183 |
| 84 | 0.0031 |  |  |  |  |  |  |  |  |  |  |
| 85 |  | 0.0051 | 0.0159 | 0.011 |  |  |  | 0.048 |  |  | 0.0041 |
| 86 | 0.02291 | 0.0196 | 0.0541 | 0.0485 |  |  |  |  |  |  | 0.0218 |
| 88 | 0.00143 | 0.0046 | 0.0135 | 0.0121 |  |  |  |  |  |  | 0.0046 |
| 90 | 0.0345 | 0.0351 | 0.0763 | 0.0594 | 0.0345 | 0.107 | 0.037 | 0.04 |  | 0.015 | 0.0338 |
| 92 | 0.0035 |  | 0.0156 | 0.0122 | 0.0052 |  |  |  |  |  |  |
| 93 |  |  |  | 0.0065 |  |  |  |  |  |  |  |
| 95 | 0.02317 | 0.0223 | 0.0595 | 0.0529 | 0.0311 | 0.087 | 0.025 | 0.03 |  |  | 0.0273 |
| 105 | 0.01167 | 0.009 | 0.0257 | 0.0181 |  |  | 0.0084 | 0.0096 |  |  | 0.0104 |
| 107 |  |  | 0.00419 |  |  |  |  |  |  |  |  |
| 110 | 0.0398 | 0.0334 | 0.0942 | 0.0762 | 0.0233 |  | 0.027 |  | 0.019 | 0.018 | 0.0342 |
| 114 |  |  | 0.00195 |  |  |  |  |  |  |  |  |
| 118 | 0.0115 | 0.022 | 0.0544 |  | 0.0182 | 0.065 | 0.0211 | 0.0259 | 0.0136 | 0.0113 | 0.0247 |
| 128 | 0.00275 | 0.0054 | 0.0113 |  |  |  |  |  |  |  |  |

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Appendix Table 3. Continued

| BZ | C-2 | C-3 | C-4 | C-5 | C-6 | C-7 | D-1 | D-2 | D-3 | D-4 | D-5 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 129 | 0.0685 | 0.0438 | 0.0748 | 0.0485 | 0.0661 | 0.122 | 0.053 | 0.06 |  | 0.023 | 0.0461 |
| 132 | 0.0079 | 0.014 | 0.0209 |  | 0.026 |  |  |  |  |  |  |
| 135 |  | 0.0175 | 0.0252 | 0.0141 |  |  | 0.02 |  |  |  |  |
| 136 |  |  | 0.0086 |  |  | 0.012 |  |  |  |  |  |
| 141 | 0.0096 |  | 0.0131 | 0.0067 |  |  |  |  |  |  | 0.0063 |
| 144 | 0.00095 | 0.0021 |  |  |  |  |  |  |  |  | 0.0026 |
| 146 |  | 0.0073 | 0.0089 | 0.0059 |  |  |  |  |  |  |  |
| 147 | 0.0471 | 0.0388 | 0.0534 |  | 0.0587 |  | 0.04 | 0.042 |  |  | 0.0339 |
| 153 | 0.0257 | 0.0402 | 0.0596 | 0.0369 | 0.06 |  | 0.046 | 0.05 |  | 0.021 |  |
| 156 | 0.0059 |  | 0.0074 |  | 0.0095 |  |  |  |  |  | 0.0043 |
| 158 | 0.00212 |  |  |  |  |  |  |  |  |  | 0.0034 |
| 164 | 0.00152 |  |  |  |  |  |  |  |  |  |  |
| 167 | 0.0018 |  |  |  |  |  |  |  |  |  | 0.0018 |
| 170 | 0.02155 |  |  |  |  |  |  |  |  |  | 0.0105 |
| 171 | 0.00219 |  | 0.0056 |  |  |  |  |  |  |  | 0.0038 |
| 172 |  |  | 0.0033 |  |  |  |  |  |  |  |  |
| 174 | 0.01627 |  | 0.0168 |  |  |  |  |  |  |  | 0.0108 |
| 175 |  |  | 0.001 |  |  |  |  |  |  |  |  |
| 176 |  |  |  |  |  |  |  |  |  |  | 0.00208 |
| 177 | 0.01119 | 0.0076 |  |  | 0.0111 |  |  |  |  |  |  |
| 178 | 0.00432 |  |  |  |  |  |  |  |  |  |  |
| 179 | 0.00744 |  | 0.00672 | 0.0042 | 0.0092 |  |  |  |  |  | 0.00503 |
| 180 | 0.0502 | 0.0251 | 0.0437 |  |  |  | 0.032 | 0.031 | 0.02 |  | 0.0282 |
| 183 | 0.00754 | 0.0067 | 0.0099 |  |  |  |  |  |  |  |  |
| 187 | 0.0276 |  | 0.0252 | 0.0155 |  |  |  |  |  |  | 0.0172 |
| 190 | 0.00272 |  |  |  |  |  |  |  |  |  |  |
| 194 | 0.0102 |  | 0.00975 |  |  |  |  |  |  |  |  |
| 195 | 0.00147 |  |  |  |  |  |  |  |  |  |  |
| 196 |  |  | 0.0042 |  |  |  |  |  |  |  |  |
| 198 | 0.0122 | 0.0066 |  |  |  |  |  |  |  |  | 0.0085 |
| 203 | 0.0082 | 0.0044 |  |  |  |  |  |  |  |  |  |
| 206 | 0.0056 |  | 0.0144 | 0.0065 |  |  |  |  |  |  |  |
| 208 |  |  | 0.0049 |  |  |  |  |  |  |  | 0.00596 |
| 209 |  | 0.0048 |  | 0.0076 | 0.0131 |  |  | 0.039 |  |  | 0.0178 |
| Total | 0.846 | 0.646 | 2.30 | 1.80 | 0.405 | 0.842 | 0.489 | 0.570 | 0.129 | 0.153 | 0.890 |

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Appendix Table 3. Continued.

| BZ | D-6 | D-7 | D-8 | D-9 | E-1 | E-2 | E-3 | E-4 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 |  |  |  |  |  |  | 0.00772 | 0.0059 |
| 2 | 0.0024 |  |  |  |  | 0.0022 |  |  |
| 3 | 0.003 |  |  |  |  |  | 0.00271 | 0.004 |
| 8 |  |  |  |  | 0.057 |  | 0.037 |  |
| 11 |  |  |  |  |  | 0.026 |  |  |
| 17 | 0.012 |  |  |  |  |  |  |  |
| 18 | 0.0261 | 0.02 |  |  |  | 0.014 | 0.0208 | 0.013 |
| 19 |  |  |  |  |  |  | 0.0077 |  |
| 20 | 0.041 | 0.0303 | 0.0103 | 0.0304 | 0.103 | 0.02 | 0.0271 | 0.0167 |
| 21 | 0.0167 | 0.0097 |  |  | 0.054 | 0.0101 | 0.0144 | 0.0081 |
| 22 | 0.0117 |  |  |  | 0.035 | 0.00719 | 0.00924 | 0.0055 |
| 25 | 0.00557 | 0.0059 |  |  |  |  |  |  |
| 26 | 0.0112 | 0.0096 |  |  | 0.0144 |  | 0.00585 | 0.0025 |
| 31 | 0.0316 | 0.0231 | 0.0072 | 0.0231 | 0.0776 | 0.0136 | 0.021 | 0.0131 |
| 32 | 0.0089 |  |  |  |  |  |  |  |
| 35 | 0.00154 |  |  |  |  |  |  |  |
| 37 | 0.015 | 0.0086 |  |  | 0.014 | 0.0049 | 0.00825 | 0.0042 |
| 40 |  | 0.0117 |  |  |  |  | 0.0078 |  |
| 44 | 0.0655 | 0.0271 |  | 0.0276 | 0.088 | 0.0208 | 0.0217 | 0.0095 |
| 45 | 0.0122 |  |  |  |  | 0.00406 | 0.0056 |  |
| 46 | 0.0028 |  |  |  |  |  |  |  |
| 48 | 0.0098 |  |  |  |  | 0.0018 |  |  |
| 49 | 0.0396 | 0.0172 |  | 0.0179 |  | 0.00559 | 0.00807 | 0.0047 |
| 50 |  |  |  |  |  |  | 0.00317 |  |
| 52 | 0.0659 | 0.0315 |  | 0.0431 | 0.034 | 0.011 | 0.0146 |  |
| 56 | 0.0289 | 0.0092 |  |  |  | 0.00315 | 0.0057 |  |
| 60 | 0.0217 | 0.0057 |  |  |  |  | 0.0044 |  |
| 61 | 0.141 | 0.0337 |  |  | 0.045 | 0.016 | 0.0225 | 0.0143 |
| 63 | 0.004 |  |  |  |  |  |  |  |
| 64 |  | 0.0103 |  |  |  | 0.00405 | 0.00634 |  |
| 66 | 0.0668 | 0.0182 |  | 0.0135 | 0.02 | 0.00748 | 0.0108 | 0.0067 |
| 67 | 0.0025 |  |  |  |  |  |  |  |
| 68 | 0.0026 |  |  |  |  |  |  |  |
| 77 | 0.0113 | 0.0029 |  |  |  | 0.00195 |  |  |
| 82 | 0.0112 |  |  |  |  |  |  |  |
| 83 | 0.0624 | 0.0134 |  |  |  | 0.00911 |  | 0.0067 |
| 84 | 0.0194 | 0.0056 |  | 0.0071 |  | 0.0021 |  |  |
| 85 |  |  |  | 0.0023 |  |  |  |  |
| 86 | 0.0648 |  |  |  |  | 0.00935 | 0.0103 | 0.0089 |
| 88 | 0.0147 |  |  |  |  |  | 0.00143 |  |
| 90 | 0.0929 | 0.0325 | 0.0232 | 0.0344 | 0.035 | 0.0169 | 0.0162 | 0.0218 |
| 92 | 0.0163 |  |  |  |  | 0.00284 |  |  |
| 95 | 0.0527 | 0.0232 | 0.0213 | 0.0336 | 0.022 | 0.0102 | 0.011 |  |
| 105 | 0.0404 |  |  | 0.0088 |  | 0.0057 |  | 0.0051 |
| 107 | 0.00626 |  |  |  |  | 0.00126 |  |  |
| 110 | 0.0996 | 0.0357 | 0.0185 | 0.0323 | 0.026 |  |  | 0.0149 |
| 118 | 0.0993 | 0.0267 | 0.0145 | 0.0246 | 0.0173 | 0.0135 |  | 0.014 |
| 123 | 0.00207 |  |  |  |  |  |  |  |
| 128 | 0.0115 |  |  |  |  | 0.0029 |  |  |
| 129 | 0.109 | 0.0496 | 0.0586 | 0.0656 | 0.049 | 0.038 | 0.0302 | 0.0365 |
| 132 | 0.0268 | 0.0142 | 0.0245 |  |  | 0.0077 | 0.0059 |  |
| 135 | 0.0313 | 0.0171 | 0.0213 | 0.0206 |  | 0.0111 | 0.00999 |  |
| 136 | 0.0092 |  |  | 0.0071 |  | 0.00282 |  |  |
| 137 | 0.0038 |  |  |  |  |  |  |  |

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Appendix Table 3. Continued.

| BZ | D-6 | D-7 | D-8 | D-9 | E-1 | E-2 | E-3 | E-4 |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 141 |  | 0.0102 |  | 0.0161 |  | 0.0062 | 0.0042 | 0.0074 |
| 144 | 0.0047 |  |  |  |  | 0.00152 |  |  |
| 146 | 0.0125 |  | 0.0098 |  |  | 0.0043 | 0.0034 |  |
| 147 | 0.0596 | 0.0367 | 0.0448 | 0.0512 | 0.036 | 0.0218 |  | 0.0248 |
| 153 | 0.0917 | 0.0433 | 0.0517 | 0.0567 | 0.044 | 0.0308 | 0.0285 | 0.0302 |
| 156 | 0.0134 |  |  |  |  | 0.00405 | 0.003 | 0.003 |
| 158 | 0.0105 |  |  |  |  | 0.0032 | 0.0019 |  |
| 164 | 0.0052 |  |  |  |  |  | 0.0017 |  |
| 167 | 0.00522 |  |  |  |  |  |  |  |
| 170 | 0.0274 |  | 0.0198 |  |  | 0.0109 | 0.00993 | 0.0094 |
| 171 | 0.0072 |  |  |  |  |  |  |  |
| 174 | 0.0223 | 0.0114 |  |  |  | 0.0078 | 0.0073 | 0.0081 |
| 176 |  |  |  |  |  | 0.0017 |  |  |
| 177 | 0.0127 | 0.0066 |  |  |  |  |  |  |
| 179 | 0.00821 |  |  |  |  |  |  |  |
| 180 | 0.0607 | 0.0265 | 0.0344 | 0.0315 |  | 0.0262 | 0.0236 |  |
| 183 | 0.0161 |  |  |  |  | 0.0059 |  |  |
| 187 | 0.0327 | 0.0164 | 0.0169 |  |  | 0.0144 | 0.0133 |  |
| 189 |  |  |  |  |  |  | 0.00083 |  |
| 190 |  |  |  |  |  | 0.0022 |  |  |
| 194 | 0.0103 | 0.0064 |  |  |  | 0.00432 |  |  |
| 195 | 0.0043 |  |  |  |  |  |  |  |
| 196 | 0.00534 |  |  |  |  | 0.0019 |  |  |
| 198 | 0.0135 | 0.0089 |  |  |  |  |  |  |
| 201 | 0.00121 |  |  |  |  |  |  |  |
| 203 | 0.0083 |  |  |  |  |  |  |  |
| 206 | 0.0108 | 0.0332 |  | 0.0171 |  |  |  |  |
| 208 | 0.005 | 0.0135 |  |  |  |  |  |  |
| 209 | 0.0245 | 0.054 |  |  |  |  | 0.00639 |  |
| Total | 1.92 | 0.760 | 0.377 | 0.565 | 0.771 | 0.455 | 0.466 | 0.299 |

A separate set of samples codes are used for the PCDD/Fs:
Appendix Table 4. PCDD/F sample code.

| sample | code |  | sample | code |
| :--- | :--- | :--- | :--- | :--- |
| 18-Mile Cr.-3/24/2008-SA | A-1 |  | Genesee R.-3/4/2008-SA | C-1 |
| 18-Mile Cr.-3/3/2008-SA | A-2 |  | Genesee R.-4/8/2008-SA | C-2 |
| 18-Mile Cr.-4/29/2008-SA | A-3 |  | Genesee R.-5/2/2007-SA | C-3 |
| 18-Mile Cr.-4/7/2008-SA | A-4 |  | Genesee R.-6/10/2008-SA | C-4 |
| 18-Mile Cr.-5/1/2007-SA | A-5 |  | Genesee R.-6/19/2008-SA | C-5 |
| 18-Mile Cr.-5/27/2008-DU | A-6 |  | Genesee R.-9/30/2008-SA | C-6 |
| 18-Mile Cr.-5/27/2008-SA | A-7 |  | Oswego R.-1/10/2008-SA | D-1 |
| 18-Mile Cr.-6/9/2008-SA | A-8 |  | Oswego R.-10/25/2007-SA | D-2 |
| Black R.-1/9/2008-SA | B-1 |  | Oswego R.-3/25/2008-SA | D-3 |
| Black R.-10/24/2007-SA | B-2 | Oswego R.-3/4/2008-SA | D-4 |  |
| Black R.-3/13/2008-SA | B-3 | Oswego R.-4/30/2008-SA | D-5 |  |
| Black R.-5/28/2008-SA | B-4 | Oswego R.-5/2/2007-SA | D-6 |  |
| Black R.-5/3/2007-SA | B-5 | Oswego R.-9/30/2008-SA | D-7 |  |
| Black R.-6/10/2008-SA | B-6 | Salmon R.-1/10/2008-SA | E-1 |  |
| Equip. Blank-2/27/2008-FB | EB-1 | Salmon R.-4/8/2008-SA | E-2 |  |
| Equip. Blank-6/11/2008-FB | EB-2 | Salmon R.-5/1/2008-SA | E-3 |  |
|  |  | Salmon R.-5/3/2007-SA | E-4 |  |

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Appendix Table 5. Site mean PCDD/F detection limits in fg/L

|  | 18-Mile <br> Cr. | Black <br> R. | Genesee <br> R. | Oswego <br> R. | Salmon <br> R. |
| ---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 15.1 | 3.5 | 5.16 | 2.46 | 2.23 |
| 2 | 9.29 | 8.03 | 5 | 2.05 | 2.86 |
| 3 | 6.02 | 3.26 | 4.59 | 2.02 | 1.91 |
| 4 | 126 | 4.14 | 4.88 | 1.88 | 2.01 |
| 5 | 6.32 | 3.11 | 4.47 | 1.85 | 1.87 |
| 6 | 20.8 | 3.37 | 6.17 | 1.99 | 2.04 |
| 7 | 23.1 | 7.58 | 7.11 | 3.64 | 3.12 |
| 8 | 40.7 | 6.18 | 20.7 | 9.83 | 18.1 |
| 9 | 69.6 | 3.79 | 5.52 | 1.95 | 8.42 |
| 10 | 16.5 | 3.37 | 5.21 | 2.19 | 1.99 |
| 11 | 6.26 | 3.58 | 4.81 | 1.82 | 1.8 |
| 12 | 626 | 9.13 | 66.7 | 16.9 | 20.2 |
| 13 | 6.56 | 3.1 | 5.16 | 1.94 | 1.99 |
| 14 | 6.82 | 3.17 | 5.65 | 2.05 | 2.12 |
| 15 | 13.9 | 209 | 363 | 136 | 85.3 |
| 16 | 16.6 | 9.38 | 7.11 | 4.43 | 5.47 |
| 17 | 8.86 | 6.21 | 7.27 | 3.63 | 3.56 |

PCDD/F congeners are designated as stated above in Table 10. Raw masses, reported by the lab as $\mathrm{pg} /$ sample, have been divided by the number of L filtered to yield concentration units ( $\mathrm{fg} / \mathrm{L}$ ).

Appendix Table 6. PCDD/F results, fg/L.

| Code | $\mathrm{A}-1$ | $\mathrm{~A}-2$ | $\mathrm{~A}-3$ | $\mathrm{~A}-4$ | $\mathrm{~A}-5$ | $\mathrm{~A}-6$ | $\mathrm{~A}-7$ | $\mathrm{~A}-8$ | $\mathrm{~B}-1$ | $\mathrm{~B}-2$ | $\mathrm{~B}-3$ | $\mathrm{~B}-4$ |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 |  |  |  |  |  |  |  |  | 8.91 |  |  |  |
| 2 | 11.8 | 116 | 7.88 | 11.5 | 15.8 | 42.3 | 33.9 | 61.7 |  | 14.1 | 2.53 | 2.5 |
| 3 | 18.6 | 141 | 14.1 | 20.9 | 24.1 | 90.4 | 69.5 | 118 | 54.1 | 22 | 4.38 | 4.61 |
| 4 | 128 |  | 105 | 171 | 186 | 726 | 584 | 1060 | 154 | 74.9 |  | 12 |
| 5 | 54.7 | 405 | 36.2 | 51.5 | 66.4 | 223 | 188 | 319 | 150 | 52.5 | 8.42 | 9.77 |
| 6 | 3370 | 25200 | 2420 | 4210 | 4820 | 19000 | 13500 | 24200 | 2810 | 2150 | 194 | 242 |
| 7 | 36700 | 300000 | 31700 | 56500 | 53000 | 247000 | 165000 | 277000 | 21300 | 24600 | 1390 | 2150 |
| 8 | 41.8 | 303 | 60.3 | 62.5 | 78 | 269 | 223 | 403 |  |  | 6.73 |  |
| 9 | 34.2 | 254 |  |  | 38.7 |  |  | 279 | 21 | 20.7 |  | 3.75 |
| 10 | 57.1 | 437 | 61.6 |  | 70.6 | 325 | 275 | 477 | 52.6 | 24.3 |  | 6.8 |
| 11 | 215 | 1720 | 174 | 305 | 322 | 1230 | 1150 | 1890 | 114 | 43.7 |  | 7.43 |
| 12 |  | 1800 |  |  | 115 |  |  |  | 67.5 | 25 |  |  |
| 13 | 38.1 | 283 | 28.7 | 44.7 | 38.1 | 227 | 203 | 179 | 56 | 17.7 | 4.04 | 5.79 |
| 14 | 4.01 |  | 2.33 | 3.68 |  | 15.7 | 12.5 | 23 |  |  |  |  |
| 15 | 1820 | 14300 | 1600 | 2850 | 2900 | 12800 | 11900 | 16400 |  | 519 |  |  |
| 16 | 78.4 | 599 | 54.2 | 99.2 | 88.8 | 416 | 402 | 500 |  |  |  |  |
| 17 | 2880 | 25400 | 2220 | 4690 | 4500 | 15600 | 14800 | 29800 | 1510 | 1630 | 88.6 | 130 |

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Appendix Table 6. Continued.

| Code | B-5 | B-6 | EB-1 | EB-2 | C-1 | C-2 | C-3 | C-4 | C-5 | C-6 | D-1 | D-2 |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 |  |  |  |  | 6.75 |  | 7.76 | 5.96 |  | 4.81 |  |  |
| 2 | 2.78 |  |  |  | 13.5 | 15.2 | 12.9 | 16.5 | 19.8 | 13.4 | 7.15 | 14.9 |
| 3 | 4.48 | 8.03 |  |  | 20.6 | 21.2 | 16.5 | 31.8 | 37.4 | 24.1 | 11.8 | 24.9 |
| 4 | 12.5 | 19.8 |  |  | 50 | 58.2 | 44.6 | 93.7 | 109 | 71.1 | 31.3 | 102 |
| 5 | 11.3 | 17.4 |  | 1400 | 59.6 | 54.9 | 49.8 | 77.8 | 94 | 49.6 | 32.2 | 76 |
| 6 | 264 | 383 | 3500 | 8300 | 1170 | 1310 | 1240 | 2170 | 2380 | 1700 | 583 | 2270 |
| 7 | 2160 | 3440 | 22300 | 82700 | 20400 | 32700 | 26000 | 29100 | 30500 | 23100 | 4640 | 21200 |
| 8 | 10.4 |  |  |  | 10.4 |  |  | 49.6 | 58.1 | 46.1 |  |  |
| 9 |  |  | 2600 |  | 11.2 | 21.2 | 12.9 | 27 | 31.1 | 29.4 | 5.88 | 20.7 |
| 10 | 4.95 | 9.61 | 2400 |  | 14.3 | 35.1 | 22.3 | 49.3 | 56.6 | 57.2 | 12.4 | 28.3 |
| 11 | 8.19 | 16.2 | 2800 | 2100 | 36.8 | 68.8 | 59.5 | 182 | 214 | 137 | 24.5 | 68.7 |
| 12 | 6.18 |  | 3400 | 1900 |  |  | 26.5 |  |  | 41 | 14 | 44.5 |
| 13 | 4.79 | 7.77 | 1600 | 1400 | 17.6 | 26.5 | 16.2 | 30.9 | 32.9 | 31.9 | 12.2 | 28 |
| 14 | 2.01 |  |  |  |  |  |  |  |  |  |  | 3.01 |
| 15 |  |  |  |  |  |  |  | 1350 | 1510 |  |  | 743 |
| 16 |  | 8.43 |  |  | 20.8 |  | 18.1 | 44 | 50.2 | 37.2 |  | 48.2 |
| 17 | 158 | 216 | 3700 | 8100 | 658 | 900 | 705 | 1740 | 1550 | 1050 | 353 | 1590 |

Appendix Table 6. Continued.

| Code | D-3 | D-4 | D-5 | D-6 | D-7 | E-1 | E-2 | E-3 | E-4 |
| ---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 |  |  | 2.27 |  |  |  |  |  |  |
| 2 | 6.84 |  | 8.86 | 9.68 | 4.42 |  | 7.18 | 3.07 | 4.14 |
| 3 | 9.01 |  | 15.3 | 14 | 7.95 | 9.48 | 10.7 | 4.06 | 4.89 |
| 4 | 18.8 | 11.6 | 55.9 | 45.9 | 32.5 | 17.7 | 21.2 | 8.52 | 10.1 |
| 5 | 22.6 | 15.7 | 41.2 | 39.2 | 17.5 | 27.1 | 26.4 | 11.8 | 14 |
| 6 | 351 | 199 | 1200 | 1090 | 798 | 274 | 342 | 131 | 193 |
| 7 | 2730 | 1510 | 11800 | 8780 | 7360 | 1930 | 2290 | 869 | 1260 |
| 8 | 10.2 |  |  | 28.4 | 15.9 |  |  |  |  |
| 9 | 7.16 | 2.85 | 15.9 | 13.2 | 7.07 |  |  |  | 5.21 |
| 10 | 9.01 |  | 15.9 | 20.4 | 10.2 | 12.5 | 15.7 | 7.01 | 7.82 |
| 11 | 17.1 | 6.58 | 38.3 | 28.2 | 13 | 28.7 | 32.2 | 13 | 7.51 |
| 12 |  |  |  | 24.4 | 9.72 | 14.7 |  |  | 6.9 |
| 13 | 9.98 | 4.8 | 19.5 | 17.9 | 9.57 | 13.3 | 14.9 | 5.97 | 5.98 |
| 14 | 3.58 |  | 2.06 | 2.17 |  |  |  |  |  |
| 15 |  |  | 442 |  |  |  |  |  |  |
| 16 | 11.3 |  | 26.4 | 17.5 |  |  |  |  | 4.45 |
| 17 | 234 | 108 | 737 | 656 | 530 | 136 | 178 | 56.3 | 95.8 |

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[^0]:    ${ }^{\text {a }}$ The WQS multiplies the concentration of each congener by a toxic equivalency factor or TEF taken from the World Health Organization (1994) consensus document and by a bioaccumulation equivalency factor (BEF). The products are summed to yield a toxic equivalency (TEQ).
    ${ }^{\mathrm{b}}$ Femtogram/L or part per quintillion - $1 \times 10 \mathrm{E}-15 \mathrm{~g} / \mathrm{L}$

